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1 **An operationalized classification of Nature Based Solutions for water-related hazards: From**
2 **theory to practice**

3 Eulalia Gómez Martín ^a, María Máñez Costa ^b & Kathleen Schwerdtner Máñez ^c

4 ^a Corresponding author. Climate Service Center Germany (GERICS), Helmholtz Center Geesthacht, Chilehaus, Eingang B
5 Fischertwiete 1, 20095 Hamburg, Germany. Email address: Eulalia.gomez@hzg.de

6 ^b Climate Service Center Germany (GERICS), Helmholtz Center Geesthacht, Chilehaus, Eingang B Fischertwiete 1, 20095
7 Hamburg, Germany. Email address: maria.manez@hzg.de

8 ^c Place.Nature.Consultancy, Am Osterberg 14, 21435 Stelle, Germany. Email address: ksmanez@gmail.com

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22 **Abstract**

23 Nature Based Solutions (NBS) are currently gaining importance in the EU policy agenda as a promising ap-
24 proach to mitigate and adapt to environmental and climate change. The main advantage of NBS over other adap-
25 tation strategies is their capability to deliver multiple benefits. They support the resilience of natural processes
26 and help in reducing adaptation costs. In this paper, we address the current gaps in the literature by providing a
27 comprehensive, easy-to-use classification scheme focussing on hydrological extreme events. The classification
28 scheme is presented as a matrix and contains a portfolio of known NBS as well as the important criteria for their
29 selection. Specifically, we have included disservices/ barriers, and the potential impacts of climate change on
30 NBS. The matrix provides decision-makers with a tool that will guide them through the first phase of the com-
31 plex process when choosing the most appropriate NBS for a specific challenge. In that way, we aim to support
32 the spread of NBS in the scientific literature as well as their practical application.

33

34 **Highlights**

- 35 • An easy-to-use classification scheme for Nature Based Solutions
- 36 • Consideration of climate change impacts on Nature Based Solutions.
- 37 • Demonstration of opportunities for Nature Based Solutions in the private sector.

38 **Key words**

39 Insurance; risk; resilience; adaptation; classification

40 **Declaration of interest: none**

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46 **1. Introduction**

47 Climate change and environmental degradation are likely to accelerate losses in European GDP by up to 77% by
48 2030 (IPCC, 2014). This has boosted the development of new approaches which highlight the role of ecosys-
49 tems in reducing the socio-economic and environmental costs of climate change. The most recent and perhaps
50 also most promising approach is the concept of Nature Based Solutions (NBS). It is based on the principle that
51 enhancing and protecting natural processes provides multiple benefits for society, thereby ensuring a sustainable
52 delivery of ecosystem services (ES) and buffering the adverse impacts of climate change. For example, restoring
53 or protecting riverine ecosystems can reduce the vulnerability of eroding riverbanks against current and project-
54 ed increases of extreme rainfall, with manifold benefits to the social-ecological system around the watershed
55 (Anbumozhi et al., 2005). Their main advantage over other adaptation strategies is their capability to deliver
56 multiple benefits. The implementation of a particular NBS may create bundles of ecosystem services, together
57 generating various social, economic and environmental co-benefits. For example, restoring or protecting coastal
58 wetlands can increase resilience against storms by acting as a barrier against natural disasters. In addition, they
59 provide multiple co-benefits, such as carbon sequestration, fish provision, job creation, or tourism (Woodward
60 and Wui, 2001; Clarkson et al., 2013).

61 A number of definitions for NBS have been formulated (e.g. Cohen-Shacham et al., 2016; Maes and Jacobs,
62 2017; Eggermont et al., 2015). In this article, we use the definition applied by the European Commission which
63 understands NBS as: “... living solutions inspired by, continuously supported by and using nature, which are
64 designed to address various societal challenges in a resource-efficient and adaptable manner and to provide sim-
65 ultaneously economic, social, and environmental benefits.¹” (Maes and Jacobs, 2017). This definition highlights

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□ According to this definition, approaches such as biomimicry should be considered as NBS since they are solutions inspired by nature. However, this approach is not contemplated in the text as many of the biomimicry solutions do not deliver multiple co-benefits as they are often designed to target one specific problem. Biomimicry technology do not necessary look for finding a balance between social, environmental and economic targets.

66 the functional role of biodiversity and ecosystem functions and processes as part of an overall adaptation strategy
67 to adapt and mitigate the impacts of climate change. For example, protecting and managing wetlands to enhance
68 their water storage capacity is considered to be a NBS. On the contrary, technical adaptation measures to
69 droughts such as water cisterns installation will not be considered a NBS as they do not use ecosystem functions
70 to deliver benefits to society. Nature Based Solutions are often used as an umbrella concept, embedding a wide
71 range of conservation and sustainability measures. Terms such as green infrastructures, ecosystem based adapta-
72 tion, ecosystem based mitigation, hybrid infrastructures, ecosystem restoration or ecosystem protection are all
73 framed under the NBS concept. For this reason, the NBS definitions found in literature are deliberately vague.
74 However, all definitions share an emphasis on the need of finding a balance between social, economic and envi-
75 ronmental targets when applying NBS, and highlight the importance of their long-term sustainability. A very
76 illustrative example are the policies promoting forests as carbon sinks that have been gaining attraction over the
77 past years. Existing international agreements and initiatives often pursue afforestation targets that are more fo-
78 cused on quantity rather than quality. These projects often use monocultures with non-native species, which can
79 produce maladaptation to climate change in the long term and negatively impact biodiversity and sustainable
80 development (Seddon et al., 2019). Despite their use of ‘nature’ to address a societal challenge, such initiatives
81 fail to find the balance between social, economic and environmental targets, and would therefore not be consid-
82 ered as NBS.

83 Implementing NBS is not only a way for adapting to environmental change, but generally supports the shift to a
84 greener economy and a more sustainable society. At the same time, it helps to reduce the costs of adaptation by
85 simply diminishing the risk in the face of uncertain events.

86 Baumgärtner and Strunz (2014) argue that the implementation of NBS increases systems capability to cope with
87 extreme hydrological events (HEE), thereby representing an insurance value against unwanted regime shifts. We
88 understand insurance value as “reflecting an ecosystem’s capacity to remain in a given regime and retain its ca-
89 pacity to deliver vital ecosystem services in the face of disturbance and change” (Baumgärtner, 2008). The NBS
90 concept, therefore, enhances the capability of social-ecological systems to cope with risks through exploiting the
91 intrinsic resilience of natural processes. This makes the concept of NBS very valuable to both public and private
92 investors that want to reduce their vulnerability to HEE. In addition, it provides an opportunity to capitalize these
93 services in Natural Assurance Schemes (NAS). When talking about NAS, we refer to strategies aiming at inter-
94 nalizing the insurance value of ecosystems with the objective of improving awareness, valuation and inclusion of

95 NBS in HEE (Denjean et al., 2017). Natural Assurance Schemes are very important in terms of re-distribution of
96 HEE risks and therefore a co-benefit of NBS implementation relevant for different economic sectors, including
97 the insurance sector.

98 Despite their great potential, the spread and standardisation of NBS in the scientific literature is limited. As a
99 result, there is little uptake and implementation of the concept by national and international decision-makers. We
100 believe that this is caused by the lack of a comprehensive, concise and easy to use classification for NBS. Similar
101 to the case of the ecosystem services classification by the Millennium Ecosystem Assessment, a simple and
102 commonly accepted NBS classification would support the transfer of the concept into adaptation and risk mitiga-
103 tion plans.

104 Existing classifications are mostly descriptive and difficult to understand and use by non-experts (WWAP, 2018,
105 European Commission, 2015, European Environmental Agency, 2015, Eggermont et al., 2015). In addition, the
106 majority of classifications do not mention undesired effects or “disservices” that may arise from malfunctioning
107 or inefficient ecosystem management (European Commission, 2013, Cohen-Shacham, 2016, Zhang et al., 2007,
108 Eggermont et al., 2015). For example, the project URBAN greenUP developed an easy-to-use catalogue of NBS
109 focussing on urban areas, but gives little attention to disservices (GreenUP, 2018). The often cited Eklipse
110 framework presents a set of indicators and assessment methods to evaluate the effectiveness of NBS projects, but
111 is unable to assess the effectiveness of NBS for disaster risk reduction (Raymond et al., 2017), or to indicate pos-
112 sible disservices. Furthermore, none of the studies looking at NBS have taken the potential effects of climate
113 change into account, although it is highly likely that climate change will have significant impacts on ecosystems
114 in general (Pecl et al., 2017), and consequently on NBS. Any decision-making on a particular NBS requires sci-
115 entifically based and customized information about the potential impacts of climate change (so-called climate
116 services). To our knowledge, there is presently no classification which gathers all information relevant for mak-
117 ing decisions on implementing NBS.

118 In response to these gaps, we present a comprehensive and easy-to-use classification scheme as a basis for as-
119 sessing and evaluating NBS under different socio-economic and climatic scenarios. Scientifically validated with
120 an extensive literature review, our paper contains three substantial contributions to the NBS concept. Firstly,
121 using the case of HEE as a conceptual focus, we propose a classification of NBS for risks associated with such
122 events. Secondly, we list the co-benefits and disservices that may potentially arise when implementing NBS.
123 And thirdly, we discuss the potential impacts of climate change on NBS, which we believe is crucial for their
124 planning and management. Our classification scheme provides decision-makers with a tool to design cost-

125 effective adaptation measures able to deliver benefits under different environmental and socio-economic scenar-
 126 ios.

127

128 2. Methodological classification framework of NBS

129 Our framework is organised as a matrix (see figure 1). Following Eggermont et al. (2015), we distinguish be-
 130 tween three types of NBS according to the level of human intervention (level of engineering required for enhanc-
 131 ing the delivery of ecosystem services): Type 1, low intervention; Type 2, medium intervention; Type 3, high
 132 intervention. While the classification has correctly been criticized for being too narrow, as well as too difficult
 133 for implementation (Potschin et al., 2016), we found its intervention-based distinction to be a useful starting
 134 point for our work.

135 An extensive literature review revealed a number of important factors that were not sufficiently considered in
 136 current classification schemes. These factors are: type of risk, area, co-benefits, disservices, impact scale, and the
 137 potential effects of climate change on NBS. For this reason, within each type of NBS, we have defined subtypes.
 138 In the vertical axis are information on the type of risk (HEE) to be considered, the area where the NBS is ap-
 139 plied, the possible co-benefits of the NBS, and a column on the possible disservices. We have additionally in-
 140 cluded impact scale (local, regional, and global), and finally the potential effects of climate change on the NBS
 141 (see appendices A).

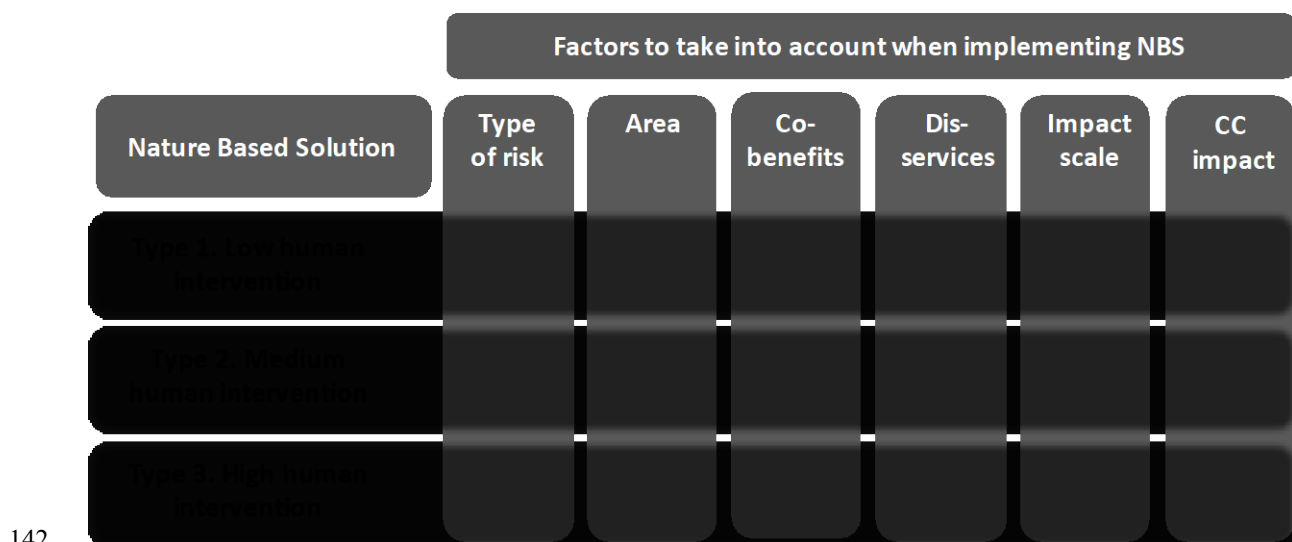


Figure 1. Synthesized Matrix for the methodological framework. The horizontal axis represents the different types of NBS according to the different levels of human intervention. The vertical axis represents key information that should be considered before NBS implementation.

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144 **2.1 Horizontal considerations of the NBS framework**

145 **2.1.1 NBS Type 1: Low human intervention**

146 Some ecosystems play a fundamental role in regulating the hydrological cycle, and in protecting against the oc-
147 currence of flood events and water scarcity (Stürck et al., 2014; Sutton-Grier et al., 2015). The type of NBS that
148 fall into this group include approaches that aim to preserve the integrity and stability of important ecological
149 functions from a group of ecosystems and habitats without an intensive management, or an intervention into the
150 system. The preservation of certain ecosystems can reduce the risks of HEE (World Bank, 2008). For example,
151 applying conservation measures in wetlands which act as barriers against storms may reduce the vulnerability of
152 coastal communities (Costanza et al., 2008). Applying measures to protect soil from wind and water erosion
153 maintains soil stability and structure and is essential to keep an optimal infiltration rate and to preserve soil fertil-
154 ity and productivity (Blanchart et al., 1999). Strategies focused on maintaining the well-functioning of ecosys-
155 tems are NBS type 1.

156 **2.1.2 NBS Type 2. Medium human intervention**

157 This type of NBS clusters all management approaches that support the enhancement of important ecosystem
158 services in a sustainable and multifunctional way. Within this group, we have included ecosystem restoration
159 approaches, and management interventions in agricultural lands, forests, river morphologies, grasslands, pastures
160 and meadows.

161 Ecosystem restoration approaches have been defined as “the process of assisting the recovery of an ecosystem
162 that has been degraded, damaged or destroyed” (SER, 2002). Frequently this degradation is the result of human
163 activities that have disturbed the ecosystem in a direct or indirect way. Some restoration projects aim to restore
164 the structure of a given ecosystem to the historic state prior to its disturbance. Others solely seek to re-establish
165 the ecological processes and functions of a given ecosystem to return to the delivery of targeted ecosystem ser-
166 vices.

167 Management interventions include a variety of measures for managing natural and man-made ecosystems. Agri-
168 cultural practices such as crop rotation can improve the fertility and structure of soil by increasing the infiltration
169 capacity. Altering deep-rooted and shallow-rooted plants can increase groundwater levels and contributes to a
170 range of other services such as pollution removal or CO₂ absorption (NWRM, 2014). Other agricultural practic-

171 es such as intercropping or green covers to protect soil from erosion can also increase the infiltration rate
172 (Zougmore et al., 2000; NWRM, 2014). Reducing soil surface exposure by maintaining an uninterrupted tree
173 canopy can also provide a number of hydrological effects such as biodiversity preservation or reduction of water
174 runoff. The reduction of exposure can be achieved with the appropriate forest management (Farley et al., 2005).
175 Another example is the establishment and maintenance of grasslands. Semi-natural grasslands buffer water flows
176 through decreasing water run-off and at the same time attenuating soil erosion. They can also decrease water
177 stress by increasing water retention capacity (Farley et al., 2005).

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179 **2.1.3 NBS Type 3: Creation of new ecosystems and hybrid solutions**

180 The last group implies higher modification of ecosystems. The creation of new ecosystems that are designed and
181 managed in a multi-purpose way are also considered type 3. It comprises green (or blue if there are aquatic eco-
182 systems involved) infrastructures which are defined as “a strategically planned network of natural and semi-
183 natural areas with other environmental features designed and managed to deliver a wide range of ecosystem ser-
184 vices” (European Commission, 2013). The combination of nature and grey infrastructure is called a hybrid solu-
185 tion and sometimes involves the creation of new ecosystems. Hybrid structures are especially useful when space
186 is limited, as is often the case in urban areas.

187 Green infrastructure and hybrid solutions are being increasingly considered in development planning. They are
188 particularly relevant in urban environments, where more than 60% of the European population lives. Measures
189 such as green roofs, permeable surface channels and rills are wide spread in cities to reduce water runoff and the
190 heat island effect (European Commission, 2013). Many of these measures have proven to be more cost-effective
191 when compared with traditional grey approaches such as dikes or levees (Liquete et al., 2016; Schäffler and
192 Swilling, 2013). For example, green alleys or tree planting have been estimated to be 3-6 times more effective in
193 managing storm-water and reducing temperatures than conventional methods. The city of Portland invested \$8
194 million in green infrastructure, saving \$250 million for hard infrastructure costs (CCAP, 2011, Liquete et al.,
195 2016; Schäffler and Swilling, 2013).

196 The creation of new ecosystems has been widely used for water retention or coastal protection measures (Piazza
197 et al., 2005). There is a number of artificial wetlands that have been constructed in seriously degraded areas with
198 water quality or drainage problems. Another example is the creation of artificial reefs to stabilize shorelines and
199 protect coasts from wave erosion and flooding (Scyphers et al., 2011). Such new habitats and ecosystems may

200 need a long time until they are fully established and able to deliver the intended ecosystem service. Until then,
201 the ecosystem can be more vulnerable to disturbances and requires protection measures.

202

203 **2.2 Vertical considerations of the NBS framework**

204 **2.2.1 Type of risk**

205 We identify several types of risks by considering their origin, spatial, and temporal dimensions. Our classifica-
206 tion covers the cases of droughts, floods, and water contamination. This is not intended to be an exhaustive list,
207 but serves as a descriptive guidance.

208 We distinguish between five different types of floods: coastal, urban, fluvial, pluvial and flash floods. Each of
209 them has different effects in terms of impacts, potential damages, and related costs. Additionally, we take into
210 account water contamination, making a distinction between organic and inorganic contamination. Within
211 droughts, we distinguish between the following types: meteorological, agricultural, hydrological, and socio-
212 economic (Liu and Kogan, 1996). In this paper, we understand a meteorological drought as the duration of the
213 dry period in reference to the normal dry period. We talk about an agricultural drought when the soil moisture is
214 insufficient and results in lack of crop growth and production. Hydrological drought is referred to periods of
215 shortfalls on surface and groundwater supply. Additionally, we consider socio-economic drought as droughts
216 associated with the mismanagement of water supply and demand (Wilhite and Glantz, 1985). Depending on the
217 type of drought, the suitability of NBS varies. For example, to reduce agricultural drought measures focused to
218 maintain soil moisture will be appropriated, while measures to reduce the impacts of socio-economic droughts
219 might include the restoration of natural vegetation and control surface flows (Sonneveld et al., 2018).

220

221 **2.2.2 Area**

222 Our approach distinguishes between three areas: rural, urban and peri-urban. Depending on the discipline, there
223 are a number of definitions used for these categories. The categorization of rural and urban spaces depends ex-
224 clusively on arbitrary delimitations, usually based on demographic components (i.e. population size), economic
225 sectoral components (i.e. percentage of population working in the primary sector) and a socio-psychological
226 component (i.e. values, attitudes, tastes and behaviours common for urban and rural areas) (Iaquinta and

227 Drescher, 2000). These components may vary depending on the area where the NBS is going to be implemented.
228 The majority of countries have their own, official definition of urban, rural and peri-urban areas.
229 For the purpose of this paper we have only considered demographic and economic sectoral components to dif-
230 ferentiate between rural and urban areas. Urban areas are densely populated areas with generally more than
231 10.000 people and with low agricultural activities. Rural areas are low densely populated areas with generally
232 less than 10.000 and with agriculture as the main economic sector. Peri-urban areas are areas currently in transi-
233 tion from strictly rural to urban.

234 Distinguishing between these categories is important, because the differences in demographic, economic and
235 socio-economic factors influence NBS. For example, as explained previously, space restrictions are especially
236 relevant for urban areas, while rural areas may have other limitations to consider, such as infrastructure access.
237 For this reason, measures that have proven to be more effective and more easily to implement in certain areas
238 (rural, urban and peri-urban), should be prioritized.

239 **2.2.3 Co-benefits**

240 One of the key aspects of NBS is their multifunctionality, their ability to provide several ecological, social, cul-
241 tural and economic benefits (Hansen and DeFries, 2007; Kabisch et al., 2017). In our classification, we distin-
242 guish between primary benefits, referring here to the intended HEE risk reduction, and secondary benefits, refer-
243 ring to additional benefits or co-benefits. For example, while the primary benefit of dunes conservation is coastal
244 flood protection, biodiversity maintenance is one of its potential co-benefits.

245 Following the Millennium Ecosystem Assessment, we classify co-benefits as provisioning, regulating, support-
246 ing and cultural ecosystem services (MEA, 2005).

247 **2.2.4 Disservices**

248 Taking limiting factors such as structural complexity, required economic investment, or available space to im-
249 plement NBS into account, as well as enabling NBS to adapt to changing conditions and disturbances is key for
250 success. If the systemic implications of NBS are not adequately considered, intended services and benefits can
251 turn into disservices. Such unintended negative side effects can potentially lead to the malfunctioning of a NBS
252 and may cause a number of undesired effects including biodiversity loss, fragmentation, change in flow patterns,
253 spread of pests and diseases, or changes in local and regional water availability (Zhang et al., 2007). For exam-
254 ple, afforestation projects to control desertification may decrease water availability if non-suitable tree species
255 are chosen. A very illustrative example is the Chinese Three Norths Shelter Forest System Project, a large-scale

256 afforestation project implemented in arid and semiarid areas to stop desertification. The project ignored key dif-
257 ferences in topography, climate and hydrology, which led to increased environmental degradation and devastat-
258 ing impacts on soil moisture, hydrology and vegetation coverage (Shixiong, 2008).

259 Identifying potential disservices is also necessary to effectively evaluate the life cycle cost of NBS, including all
260 cost associated with designing, building and maintaining a functioning NBS. The vast majority of NBS require
261 ongoing management. For example, parts of urban nature are maintained through trimming, irrigation or collect-
262 ing leaf litter. Lack of funding or inefficient planning causes management failures, potentially resulting in de-
263 creasing ecosystem services delivery, the loss of social acceptance due to accidents, unpleasant views, damages
264 in infrastructures, and several other issues, with impacts on costs. For example, if urban trees are not well man-
265 aged, infrastructures such as pavements can be damaged by tree roots of fallen limbs, requiring costly repairs.
266 The relationship between the different ecosystem components and socio-economic factors is complex. For this
267 reason, being aware of the potential disservices in the prior stages of NBS implementation may help to identify
268 factors that need to be considered in the design and evaluation the NBS. Consequently, our classification also
269 includes a column in which the most common disservices and socio-economic factors limiting the success of
270 NBS implementation are listed. This list is designed to be used in the prior stages of the decision-making pro-
271 cess. It only gives a general idea of the potential disservices that may arise if factors such as budget, maintenance
272 cost or climate change impact are not considered. The table can be used to narrow down the list of potential dis-
273 services as a check box exercise.

274

275 **2.2.5 Impact scale**

276 The co-benefits delivered by NBS can have impacts at different spatial scales. For example, a reforestation pro-
277 ject might have a primary impact at the local scale reducing soil erosion in hillsides and at the same time a global
278 impact by capturing CO₂. However, to facilitate management decisions we only consider, for each NBS, the
279 scale of the impacts that the primary benefit may have. The definition of the scales depends on the particular
280 context and should be clearly stated when assessing the impact of NBS. In our framework, we use the scale pro-
281 vided in the Eclipse framework which considers the impacts of NBS at the mesoscale (regional, metropolitan,
282 urban) and microscale (neighbourhood/street, building) (Raymond et al., 2017).

283 **2.2.6 Climate change impacts**

284 The impacts of climate change on NBS need special consideration. The latest IPCC report makes specific refer-
285 ence to the impacts of climate change on ecosystem functions due to increasing and/or decreasing climate varia-
286 bility, increase in mean temperature, change in precipitation patterns, increase of extreme events, or sea level
287 raise (IPCC, 2018). At the same time, an increase in the frequency and intensity of HEE may overwhelm the
288 capability of NBS to cope with these risks. Changes in rain patterns may also affect water and energy demand
289 exacerbating social and cultural divisions.

290 Potential impacts of climate change on NBS have not been adequately addressed in the literature so far (Seddon
291 2019). However, such considerations are of utmost importance to make sure that the effects of NBS will last,
292 especially in the long term. There are two relevant aspects: Firstly, the impacts of climate change on the perfor-
293 mance of NBS and their effectiveness. And secondly, the costs that are related to dealing with and adapting to
294 those impacts.

295 The effectiveness of NBS to cope with risk can be influenced in many ways. For example, higher-intensity of
296 rain or flood events may saturate the capability of coastal habitats or green infrastructure to deal with these risks.
297 Changes in the mean temperature, rain patterns, species distribution, fire patterns or HEE have important impli-
298 cations for practices such as ecological restoration. Future biophysical conditions arising as a result of climate
299 change as well as other factors such as habitat fragmentation or land use change create novel environmental con-
300 ditions never experienced in ecosystems before (Harris et al., 2006; Tilman and Lehman, 2001). For example, a
301 number of plants are shifting their distribution ranges to higher elevations and latitudes (Chen et al., 2011). In
302 some cases, this may challenge the capability of ecosystems to adapt and thus to deliver certain ES. Taking cli-
303 mate change into account is also key in afforestation projects, since droughts, fires, pests and diseases may de-
304 termine long-term effectiveness of the project (Zhu et al., 2011).

305 Changes in species distribution range can even have implications for human health as human disease vectors
306 such as mosquitoes are also temperature dependent. Consequently, changes in climate could lead to changes in
307 the dynamics of diseases transmission (Afrane et al., 2012; Martens et al., 1999). This should be taken into ac-
308 count when it comes to the design of NBS. Areas with high risk of being affected by this phenomenon should
309 pay more attention to the design and establishment of blue infrastructure since habitats with stagnant water play
310 an important role in mosquitoes' life cycle.

311 The resources and cost of implementing NBS may increase in drier and warmer environments. For example, the
312 maintenance of urban green infrastructures through irrigation can increase water use, potentially consuming
313 more water than they infiltrate. The effectiveness of NBS can also be threatened by future development of dis-

314 services arising from new climatic conditions. For example, blue infrastructures may be more susceptible to algal
315 blooms with increasing sea surface temperatures.

316 Considering the potential impacts of climate change in the first stages of NBS development is necessary for iden-
317 tifying factors that influence the delivery of ES and disservices associated with NBS as well as the capability of
318 NBS to resist, recover or adapt to future conditions. This can help to design connected, heterogeneous and eco-
319 logically diverse NBS able to adapt to new climatic conditions (Seddon et al., 2019).

320 **2.2.7. An example on how to use the methodological framework**

321 In 2011 a climate adaptation plan was adopted by the city Council of Copenhagen in order to address the enor-
322 mous challenges that the ongoing and future rainfall events are causing in the city. The Copenhagen Climate
323 Adaptation Plan (CCAP) sets the framework for the implementation of climate adaptive measures in the City
324 Administration area. As part of the CCAP, the Cloudburst Management Plan was developed to reduce the eco-
325 nomic and societal problems caused by extreme flood events. The challenges arising from extreme rainfall could
326 not be solved by a single initiative, such as upgrading the sewage system. For this reason, combined and coordi-
327 nated actions had to be implemented. The development of projects promoting blue and green infrastructures is
328 therefore a core aspect of the management plan (City of Copenhagen, 2012). In this section, we use the Copen-
329 hagen case to illustrate how our framework could be used in the first-stages of NBS implementation in an urban
330 area with extreme rainfall problems.

331 The first step of the decision-making process is to identify the type of risk that the system is facing. In this case,
332 the city of Copenhagen faces urban and pluvial flooding. At this stage, it is important to evaluate the probability
333 of flooding vulnerability of the society in flood-prone areas.

334 Our table (see appendices a) lists a number of measures that can be applied to reduce flooding in urban areas,
335 ranging from restoration or conservation of upstream floodplain areas to the installation of hybrid infrastructures
336 such as vegetated swales, filter strips or tree pits.

337 Given that NBS are site-specific, the success of any measure is tightly linked with the environmental and socio-
338 economic conditions of the area in which they will be applied. To effectively evaluate the potential co-benefits
339 and disservices which a NBS can produce, it is necessary to consider hydrological, climatic and socio-economic
340 studies. The table provides a description with the most commonly delivered co-benefits and disservices of each
341 NBS. This can serve as a fist analysis of the best approach that should be considered for a deeper assessment.

342 For example, pollution and heat waves have been identified as increasing problems in the city of Copenhagen.
343 This means that NBS delivering co-benefits related to climate regulation are highly likely to be more cost-
344 effective, because they address several problems at the same time. Stream bed re-naturalization is usually a suit-
345 able measure to tackle urban, fluvial and pluvial flooding. However, the high- economic investment and the re-
346 quired ongoing management cost may be a barrier to the successful implementation of this NBS . In addition,
347 this measure is only suitable if there is enough space. Other low-cost measures such as rain gardens, infiltration
348 tranches or vegetated filter strips may be more adequate if space and budget are limiting factors.

349 Once the range of available alternatives have been identified, it is important to assess their long-term effective-
350 ness, also under climate change projections. Our classification provides a first overview of the potential impacts
351 that climate change may have on NBS. For example, if green roofs have been identified as an effective NBS,
352 changes in rain patterns and species distribution should be taken into account in order to choose appropriate spe-
353 cies or considering an increase of water demand.

354

355 **3. Discussion**

356 Nature Based Solutions are thought to be a promising approach for mitigating and adapting to environmental and
357 climate change. However, their spread in the scientific literature as well as their practical implementation is cur-
358 rently hampered by the lack of a comprehensive, concise and easy to use classification. Although several classi-
359 fication schemes have been developed, we found that they have three important shortcomings. Firstly, they are
360 often descriptive, which makes both their understanding by non-experts as well as an easy applicability challeng-
361 ing (WWAP, 2018, European Commission, 2015). Secondly, they do neither account for undesired effects (the
362 so-called “disservices”), nor do they explicitly take implementation barriers into account. And thirdly, none of
363 the existing classification schemes does take the potential effects of climate change into account. In light of cur-
364 rent predications on the impacts of climate change on ecosystems and the services which they deliver, we see
365 this shortcoming as a major limitation with respect to the long-term sustainability of NBS.

366 We have addressed all three shortcomings by developing a classification scheme which is comprehensive and
367 clearly arranged. It includes potential disservices/ barriers, and specifically addresses the potential effects of cli-
368 mate change. The classification scheme is organized as a matrix, and provides a suitable and easy-to-use tool
369 which we hope will both be taken up by the scientific literature as well as be useful for decision-making.

370 It is remarkable that despite the importance of NBS in mitigating and adapting to climate change, there are al-
371 most no studies assessing their effectiveness in a climate change context. Like other ecosystems, NBS will also
372 be affected by the impacts of a changing climate. Our classification scheme provides the criteria needed for de-
373 veloping climate services as an important precondition for NBS implementation. Climate services are scientifi-
374 cally based and customised information about the potential impacts of climate change on a particular system
375 (Hewitt et al., 2012). Consequently, they need to be taken into account when choosing a NBS, to avoid increas-
376 ing costs in the future or even a potential failure.

377 Our classification scheme does have two potential limitations which we will shortly discuss here. The first is
378 some ambiguity in differentiating types of NBS based on the level of intervention required. However, this dis-
379 tinction is not always entirely clear. For example, the conservation of certain ecosystems such as semi-natural
380 grassland or some types of forests (considered as type 1) requires ongoing management. Consequently, these
381 NBS could also be included in type 2 (medium human intervention). Similar examples can be found for type 2
382 and type 3 strategies. For instance, some restoration approaches can be seen as NBS type 3 as they use grey in-
383 frastructures to recover ecosystem functions. A prime example is the use of concrete blocks to allow the estab-
384 lishment of marine life and encourage the growth of new reefs.

385 A second potential limitation is that our framework does not consider the cost of NBS implementation. We as-
386 sume that in general the management and maintenance cost increase from type 1 to type 3. If the level of engi-
387 neering or management is high, the cost of maintaining the well-functioning such type 3 NBS are also likely to
388 be high, given that such manufactured Nature Based Solutions lack the self-regulation of a purely ‘natural’ eco-
389 system. In any case, decision-making on what NSB shall be implemented will always require a careful consider-
390 ation of costs, for example in form of a cost-effectiveness analysis.

391 The primary aim of NBS is the delivery of ecosystem services, which is a basis for obtaining ecosystem benefits
392 (Schwerdtner Máñez et al. 2014). Generally speaking, all types of NBS 1 and 2 can be expected to deliver a high
393 amount of ES. These services are related to the inherent functioning of the ecosystem. While NBS type 3 might
394 deliver fewer ecosystem services, the fact that they are engineered also means that they can be designed to deliv-
395 ery specific services. As a result, they might be more effective in solving particular problems. Such engineered
396 systems may better fit into environments that do not allow for the establishment of “natural systems”, for exam-
397 ple, because of space restrictions. Hence, type 3 NBS are often found in urban areas, where they serve a particu-
398 lar aim, such as preventing urban floods.

399 The proposed classification scheme summarizes available options for NBS to HEE management. It does not only
400 intend to potentially increase the adoption of NBS measures, but also aims to raise awareness about the added
401 value of planning with NAS, namely to support the capability of ecosystems in reducing the negative effects of
402 HEE through enhancing their insurance value.

403 Proactive involvement at all societal levels is needed to enhance ecosystem resilience, first in order to analyse
404 the risks, and second to find leverage points for NBS implementation. The need of societal involvement goes far
405 beyond implementing NBS, as it is also connected to insurance companies as “redistributors” of risk. Insurance
406 companies have had important roles in the past, for example by supporting the establishment of fire departments
407 in previous times to reduce the impacts of fire and the possible losses (Chester, 2018). It is time now for them to
408 support the implementation of NBS, considering their implicit insurance value, to reduce the risks to HEE. We
409 believe that our approach can facilitate this process.

410

411 **4. Conclusions**

412 Our classification scheme is a tool intended to support the spread and implementation of Nature Based Solutions
413 as efficient measures to mitigate and adapt to environmental and climate change. Using the example of Hydro-
414 logical Extreme Events (HEE) as a conceptual focus, we provide a portfolio of NBS which have proven to be
415 effective in Disaster Risk Reduction (DRR) and climate change adaptation. Most importantly, we introduce the
416 relevant criteria for supporting the complex decision-making processes for NBS. This provides decision-makers
417 with an easy-to-use tool for NBS implementation. This is important not only for public bodies and decision-
418 makers, but also for the private sector, including insurance companies. Given that insurance companies have an
419 important role as “risk redistributors”, and considering the inherent insurance value of NBS, we believe it is now
420 time for insurance companies to get involved in NBS.

421

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426

427 **References**

428

429 Afrane, Y., Githeko, A., Yan, G., 2012. The ecology of Anopheles mosquitoes under climate change: Case stud-
 430 ies from the effects of deforestation in East African highlands, *Annals of the New York Academy of Sci-*
 431 *ences*. <https://doi.org/10.1111/j.1749-6632.2011.06432.x>

432 Anbumozhi, V., Radhakrishnan, J., Yamaji, E., 2005. Impact of riparian buffer zones on water quality and asso-
 433 ciated management considerations. *Ecol. Eng.* 24, 517–523.
 434 <https://doi.org/https://doi.org/10.1016/j.ecoleng.2004.01.007>

435 Baumgärtner, S., 2008. The insurance value of biodiversity in the provision of ecosystem services. *Nat. Resour.*
 436 *Model.* 20, 87–127. <https://doi.org/10.1111/j.1939-7445.2007.tb00202.x>

437 Baumgärtner, S., Strunz, S., 2014. The economic insurance value of ecosystem resilience. *Ecol. Econ.* 101, 21–
 438 32. <https://doi.org/https://doi.org/10.1016/j.ecolecon.2014.02.012>

439 Blanchart, E., Alain, A., Alegre, J., Duboisset, A., Villenave, C., Pashanasi, B., Lavelle, P., Brussaard, L., 1999.
 440 Effects of earthworms on soil structure and physical properties, *Earthworm Management in Tropical*
 441 *Agroecosystems*.

442 Chen, I.-C., Hill, J.K., Ohlemüller, R., Roy, D.B., Thomas, C.D., 2011. Rapid Range Shifts of Species Associat-
 443 ed with High Levels of Climate Warming. *Science* (80). 333, 1024 LP-1026.

444 Chester, M., 2018. Insurers, bond rating companies, and local governments as proponents of natural fortifica-
 445 tions in vulnerable communities [Online] Available at:
 446 [https://www.thenatureofcities.com/2015/09/29/what-is-the-insurance-value-of-urban-ecosystems-and-](https://www.thenatureofcities.com/2015/09/29/what-is-the-insurance-value-of-urban-ecosystems-and-their-services/)
 447 [their-services/](https://www.thenatureofcities.com/2015/09/29/what-is-the-insurance-value-of-urban-ecosystems-and-their-services/) [Accessed July 2018]

448 City of Copenhagen, 2012. Cloudburst management plan. Copenhagen.

449 City of GreenUP, U., 2018. New Strategy for Re-Naturing Cities through Nature-Based Solutions – NBS Cata-
 450 logue.

451 Cohen-Shacham, G., Walters, C.J., Maginnis, S., 2016. Nature-based Solutions to address global societal chal-
 452 lenges. Switzerland. <https://doi.org/http://dx.doi.org/10.2305/IUCN.CH.2016.13.en>

- 453 Costanza, R., Pérez-Maqueo, O., Martinez, M.L., Sutton, P., Anderson, S.J., Mulder, K., 2008. The Value of
454 Coastal Wetlands for Hurricane Protection. *AMBIO A J. Hum. Environ.* 37, 241–248.
455 [https://doi.org/10.1579/0044-7447\(2008\)37\[241:TVOCWF\]2.0.CO;2](https://doi.org/10.1579/0044-7447(2008)37[241:TVOCWF]2.0.CO;2)
- 456 Denjean, B., Denjean, B., Altamirano, M.A., Graveline, N., Giordano, R., Van der Keur, P., Moncoulon, D.,
457 Weinberg, J., Máñez Costa, M., Kozinc, Z., Mulligan, M., Pengal, P., Matthews, J., van Cauwenbergh, N.,
458 López Gunn, E., Bresch, D.N., Denjean, B., 2017. Natural Assurance Scheme: A level playing field
459 framework for Green-Grey infrastructure development. *Environ. Res.* 159, 24–38.
460 <https://doi.org/10.1016/j.envres.2017.07.006>
- 461 Eggermont Hilde, Balian, E., Azevedo, J.M.N., Beumer, V., Brodin, T., Claudet, J., Fady, B., Grube, M., Keune,
462 H., Lamarque, P., Reuter, K., Smith, M., van Ham, C., Weisser, W.W., Le Roux, X., 2015. Nature-based
463 Solutions: New Influence for Environmental Management and Research in Europe. *GAIA - Ecol. Perspect.*
464 *Sci. Soc.* 24, 243–248(6). <https://doi.org/https://doi.org/10.14512/gaia.24.4.9>
- 465 European Commission., 2013. Communication from the Commission to the European parliament, the council,
466 the European economic and social committee and the committee of the regions. Green Infrastructure (GI)
467 — Enhancing Europe’s Natural Capital. Brussels.
- 468 European Commission., 2015. Towards an EU Research and Innovation policy agenda for Nature-Based Solu-
469 tions & Re-Naturing Cities Final Report of the Horizon 2020 Expert Group on “Nature-Based Solutions
470 and Re-Naturing Cities.” Brussels. <https://doi.org/10.2777/765301>
- 471 European Environmental Agency., 2015. Exploring nature-based solutions. The role of green infrastructure in
472 mitigating the impacts of weather-and climate change-related natural hazards. Luxembourg.
473 <https://doi.org/10.2800/946387>
- 474 Faivre, N., Fritz, M., Freitas, T., de Boissezon, B., Vandewoestijne, S., 2017. Nature-Based Solutions in the EU:
475 Innovating with nature to address social, economic and environmental challenges. *Environ. Res.* 159, 509–
476 518. <https://doi.org/https://doi.org/10.1016/j.envres.2017.08.032>
- 477 Farley, K.A., Jobbágy, E.G., Jackson, R.B., 2005. Effects of afforestation on water yield: a global synthesis with
478 implications for policy. *Glob. Chang. Biol.* 11, 1565–1576. [https://doi.org/10.1111/j.1365-](https://doi.org/10.1111/j.1365-2486.2005.01011.x)
479 [2486.2005.01011.x](https://doi.org/10.1111/j.1365-2486.2005.01011.x)

- 480 GreenUP, U., 2018. New Strategy for Re-Naturing Cities through Nature-Based Solutions – NBS Catalogue.
- 481 Hansen, A.J., DeFries, R., 2007. Ecological mechanisms linking protected areas to surrounding lands. *Ecol.*
482 *Appl.* 17, 974–988. <https://doi.org/10.1890/05-1098>
- 483 Harris, J., J. Hobbs, R., Higgs, E., Aronson, J., 2006. Ecological Restoration and Global Climate Change, *Resto-*
484 *ration Ecology - RESTOR ECOL.* <https://doi.org/10.1111/j.1526-100X.2006.00136.x>
- 485 Hewitt, C., Mason, S., Walland, D., 2012. The Global Framework for Climate Services. *Nat. Clim. Chang.* 2,
486 831.
- 487 Iaquina, D., Drescher, A.W., 2000. Defining the peri-urban: Rural-urban linkages and institutional connections,
488 *Land Reform, Land Settlement and Cooperatives.*
- 489 IPCC., 2014. *Climate Change 2014: Synthesis Report. Contribution of Working Groups I, II and III to the Fifth*
490 *Assessment Report of the Intergovernmental Panel on Climate Change [Core Writing Team, R.K. Pachauri*
491 *and L.A. Meyer (eds.)]. IPCC, Geneva, Switzerland, 151 pp*
- 492 IPCC., 2018. *Global warming of 1.5°C. An IPCC Special Report on the impacts of global warming of 1.5°C*
493 *above pre-industrial levels and related global greenhouse gas emission pathways, in the context of*
494 *strengthening the global response to the threat of climate change. , eds. V. Masson-Delmotte, P. Zhai, H.*
495 *O. Pörtner, D. Roberts, J. Skea, P. R. Shukla, et al. Geneva, Switzerland: World Meteorological Organiza-*
496 *tion.*
- 497 Kabisch, N., Korn, H., Stadler, J., Bonn, A., 2017. *Nature-Based Solutions to Climate Change Adaptation in*
498 *Urban Areas—Linkages Between Science, Policy and Practice* BT - *Nature-Based Solutions to Climate*
499 *Change Adaptation in Urban Areas: Linkages between Science, Policy and Practice*, in: Kabisch, N., Korn,
500 H., Stadler, J., Bonn, A. (Eds.), . Springer International Publishing, Cham, pp. 1–11.
501 https://doi.org/10.1007/978-3-319-56091-5_1
- 502 Liqueste, C., Udias, A., Conte, G., Grizzetti, B., Masi, F., 2016. Integrated valuation of a nature-based solution
503 for water pollution control. Highlighting hidden benefits. *Ecosyst. Serv.* 22, 392–401.
504 <https://doi.org/10.1016/j.ecoser.2016.09.011>
- 505 LIU, W.T., KOGAN, F.N., 1996. Monitoring regional drought using the Vegetation Condition Index. *Int. J. Re-*

- 506 mote Sens. 17, 2761–2782. <https://doi.org/10.1080/01431169608949106>
- 507 Maes, J., Jacobs, S., 2017. Nature-Based Solutions for Europe’s Sustainable Development. *Conserv. Lett.* 10,
508 121–124. <https://doi.org/10.1111/conl.12216>
- 509 Martens, P., Kovats, R.S., Nijhof, S., de Vries, P., Livermore, M.T.J., Bradley, D.J., Cox, J., McMichael, A.J.,
510 1999. Climate change and future populations at risk of malaria. *Glob. Environ. Chang.* 9, S89–S107.
511 [https://doi.org/https://doi.org/10.1016/S0959-3780\(99\)00020-5](https://doi.org/https://doi.org/10.1016/S0959-3780(99)00020-5)
- 512 Munang, R., Thiaw, I., Alverson, K., Liu, J., Han, Z., 2013. The role of ecosystem services in climate change
513 adaptation and disaster risk reduction. *Curr. Opin. Environ. Sustain.* 5, 47–52.
514 <https://doi.org/https://doi.org/10.1016/j.cosust.2013.02.002>
- 515 NWRM, 2014. Natural Water Retention Measures [Online] Available at: <http://nwrn.eu/> [Accessed May 2018]
- 516 Pecl, G.T., Araújo, M.B., Bell, J.D., Blanchard, J., Bonebrake, T.C., Chen, I.-C., Clark, T.D., Colwell, R.K.,
517 Danielsen, F., Evengård, B., Falconi, L., Ferrier, S., Frusher, S., Garcia, R.A., Griffis, R.B., Hobday, A.J.,
518 Janion-Scheepers, C., Jarzyna, M.A., Jennings, S., Lenoir, J., Linnetved, H.I., Martin, V.Y., McCormack,
519 P.C., McDonald, J., Mitchell, N.J., Mustonen, T., Pandolfi, J.M., Pettoelli, N., Popova, E., Robinson,
520 S.A., Scheffers, B.R., Shaw, J.D., Sorte, C.J.B., Strugnell, J.M., Sunday, J.M., Tuanmu, M.-N., Vergés, A.,
521 Villanueva, C., Wernberg, T., Wapstra, E., Williams, S.E., 2017. Biodiversity redistribution under climate
522 change: Impacts on ecosystems and human well-being. *Science* (80-.). 355.
- 523 Piazza, B.P., Banks, P.D., La Peyre, M.K., 2005. The Potential for Created Oyster Shell Reefs as a Sustainable
524 Shoreline Protection Strategy in Louisiana. *Restor. Ecol.* 13, 499–506. [https://doi.org/10.1111/j.1526-](https://doi.org/10.1111/j.1526-100X.2005.00062.x)
525 [100X.2005.00062.x](https://doi.org/10.1111/j.1526-100X.2005.00062.x)
- 526 Potschin, M., Kretsch, C., Haines-Young, R., E. Furman, Berry, P., Baró, F., 2016. Nature-based solutions. In:
527 Potschin, M. and K. Jax (eds): *OpenNESS Ecosystem Services Reference Book*. EC FP7 Grant Agreement
528 no. 308428. Available via: www.openness-project.eu/library/reference-book
- 529 Raymond, C., Berry, P., Breil, M., Nita, M., Kabisch, N., de Bel, M., Enzi, V., Frantzeskaki, N., Geneletti, D.,
530 Cardinaletti, M., Lovinger, L., Basnou, C., Monteiro, A., Robrecht, H., Sgrigna, G., Munari, L. and
531 Calfapietra, C., 2017. An Impact Evaluation Framework to Support Planning and Evaluation of Nature-
532 based Solutions Projects. Report prepared by the EKLIPSE Expert Working Group on Nature-based Solu-

- 533 tions to Promote Climate Resilience in Urban Areas. Wallingford, United Kingdom.
- 534 Raymond, C., Berry, P., Breil, M., Nita, M., Kabisch, N., de Bel, M., Enzi, V., Frantzeskaki, N., Geneletti, D.,
535 Cardinaletti, M., Lovinger, L., Basnou, C., Monteiro, A., Robrecht, H., Sgrigna, G., Munari, L. and Calfa-
536 pietra, C., 2017. An Impact Evaluation Framework to Support Planning and Evaluation of Nature-based
537 Solutions Projects. Report prepared by the EKLIPSE Expert Working Group on Nature-based Solutions to
538 Promote Climate Resilience in Urban Areas. Wallingford, United Kingdom.
- 539 Schäffler, A., Swilling, M., 2013. Valuing green infrastructure in an urban environment under pressure — The
540 Johannesburg case. *Ecol. Econ.* 86, 246–257.
541 <https://doi.org/https://doi.org/10.1016/j.ecolecon.2012.05.008>
- 542 Schwerdtner Máñez, K., Krause, G., Glaser, M. 2014. The Gordian knot of mangrove conservation: distangling
543 the role of scale, services, and benefits. *Global Environmental Change* 28, 120-128.
- 544 Scyphers, S.B., Powers, S.P., Heck Jr, K.L., Byron, D., 2011. Oyster Reefs as Natural Breakwaters Mitigate
545 Shoreline Loss and Facilitate Fisheries. *PLoS One* 6, e22396.
- 546 Seddon, N., Turner, B., Berry, P., Chausson, A., Girardin, C.A.J., 2019. Grounding nature-based climate solu-
547 tions in sound biodiversity science. *Nat. Clim. Chang.* 9, 84–87. [https://doi.org/10.1038/s41558-019-0405-](https://doi.org/10.1038/s41558-019-0405-0)
548 [0](https://doi.org/10.1038/s41558-019-0405-0)
- 549 SER, P., 2002. SER Primer on Ecological Restoration [WWW Document]. 2002. URL www.ser.org
- 550 Shixiong, C., 2008. Why large-scale afforestation efforts in China have failed to solve the desertification prob-
551 lem.
- 552 Sonneveld, B.G.J.S. Merbis, M.D. Alfara, A. & Ünver, O. and Arnal, M.A., 2018. Nature-Based Solutions for
553 agricultural water management and food security. *FAO Land and Water Discussion Paper no. 12*. Rome,
554 FAO. 66 pp.
- 555 Stürck, J., Poortinga, A., Verburg, P.H., 2014. Mapping ecosystem services: The supply and demand of flood
556 regulation services in Europe. *Ecol. Indic.* 38, 198–211.
557 <https://doi.org/https://doi.org/10.1016/j.ecolind.2013.11.010>
- 558 Sutton-Grier, A.E., Wowk, K., Bamford, H., 2015. Future of our coasts: The potential for natural and hybrid

- 559 infrastructure to enhance the resilience of our coastal communities, economies and ecosystems. *Environ.*
560 *Sci. Policy* 51, 137–148. [https://doi.org/https://doi.org/10.1016/j.envsci.2015.04.006](https://doi.org/10.1016/j.envsci.2015.04.006)
- 561 Tilman, D., Lehman, C., 2001. Human-caused environmental change: Impacts on plant diversity and evolution.
562 *Proc. Natl. Acad. Sci.* 98, 5433 LP-5440.
- 563 Wilhite, D., Glantz, M., 1985. Understanding: the Drought Phenomenon: The Role of Definitions. *Water Int.* -
564 *WATER INT* 10, 111–120. <https://doi.org/10.1080/02508068508686328>
- 565 Woodward, R.T., Wui, Y.-S., 2001. The economic value of wetland services: a meta-analysis. *Ecol. Econ.* 37,
566 257.
- 567 World Bank., 2008. Biodiversity, Climate Change, and Adaptation: Nature-based Solutions from the World
568 Bank Portfolio. Washington, DC.
- 569 WWAP (United Nations World Water Assessment Programme)/UN-Water, 2018. The United Nations World
570 Water Development Report 2018: Nature-Based Solutions for Water. Paris.
- 571 Zhang, W., Ricketts, T.H., Kremen, C., Carney, K., Swinton, S.M., 2007. Ecosystem services and dis-services to
572 agriculture. *Ecol. Econ.* 64, 253–260. [https://doi.org/https://doi.org/10.1016/j.ecolecon.2007.02.024](https://doi.org/10.1016/j.ecolecon.2007.02.024)
- 573 Zhu, K., Woodall, C.W., Clark, J.S., 2011. Failure to migrate: lack of tree range expansion in response to climate
574 change. *Glob. Chang. Biol.* 18, 1042–1052. <https://doi.org/10.1111/j.1365-2486.2011.02571.x>
- 575 Zougmore, R., Kambou, F.N., Ouattara, K., Guillobez, S., 2000. Sorghum-cowpea Intercropping: An Effective
576 Technique Against Runoff and Soil Erosion in the Sahel (Saria, Burkina Faso). *Arid Soil Res. Rehabil.* 14,
577 329–342. <https://doi.org/10.1080/08903060050136441>
- 578
- 579

585 **Appendices A References**

- 586 [1] Anbumozhi, V., Radhakrishnan, J., Yamaji, E., 2005. Impact of riparian buffer zones on water quality
587 and associated management considerations. *Ecol. Eng.* 24, 517–523.
588 <https://doi.org/https://doi.org/10.1016/j.ecoleng.2004.01.007>
- 589 [2] Baker, A.C., Starger, C.J., McClanahan, T.R., Glynn, P.W., 2004. Corals' adaptive response to
590 climate change. *Nature* 430, 741.
- 591 [3] Barbier, E.B., Koch, E.W., Silliman, B.R., Hacker, S.D., Wolanski, E., Primavera, J., Granek, E.F., Po-
592 lasky, S., Aswani, S., Cramer, L.A., Stoms, D.M., Kennedy, C.J., Bael, D., Kappel, C. V, Perillo,
593 G.M.E., Reed, D.J., 2008. Coastal Ecosystem-Based Management with Nonlinear Ecological Functions
594 and Values. *Science* (80-.). 319, 321 LP-323. <https://doi.org/10.1126/science.1150349>
- 595 [4] Bautista, S., Bellot, J., Vallejo, V.R., 1996. Mulching treatment for postfire soil conservation in a semi-
596 arid ecosystem. *Arid Soil Res. Rehabil.* 10, 235–242. <https://doi.org/10.1080/15324989609381438>
- 597 [5] Bizoza, A.R., de Graaff, J., 2012. Financial cost–benefit analysis of bench terraces in Rwanda. *L. De-*
598 *grad. Dev.* 23, 103–115. <https://doi.org/10.1002/ldr.1051>
- 599 [6] Brix, H., 1997. Do macrophytes play a role in constructed treatment wetlands? *Water Sci. Technol.* 35,
600 11–17. [https://doi.org/https://doi.org/10.1016/S0273-1223\(97\)00047-4](https://doi.org/https://doi.org/10.1016/S0273-1223(97)00047-4)
- 601 [7] Bullock, D.G., 1992. Crop rotation. *CRC. Crit. Rev. Plant Sci.* 11, 309–326.
602 <https://doi.org/10.1080/07352689209382349>
- 603 [8] Ciriacy-Wantrup, S. V., 1947. Capital Returns from Soil-Conservation Practices. *J. Farm Econ.* 29,
604 1181–1196. <https://doi.org/10.2307/1232747>
- 605 [9] Collins, S.L., Knapp, A.K., Briggs, J.M., Blair, J.M., Steinauer, E.M., 1998. Modulation of Diversity by
606 Grazing and Mowing in Native Tallgrass Prairie. *Science* (80-.). 280, 745 LP-747.
607 <https://doi.org/10.1126/science.280.5364.745>
- 608 [10] European Commission., 2015. Towards an EU Research and Innovation policy agenda for Nature-
609 Based Solutions & Re-Naturing Cities Final Report of the Horizon 2020 Expert Group on “Nature-
610 Based Solutions and Re-Naturing Cities.” Brussels. <https://doi.org/10.2777/765301>

- 611 [11] Costanza, R., Farber, S.C., Maxwell, J., 1989. The valuation and management of wetland ecosystems.
612 Ecol. Econ. 1, 335.
- 613 [12] Costanza, R., Pérez-Maqueo, O., Martínez, M.L., Sutton, P., Anderson, S.J., Mulder, K., 2008. The Va-
614 lue of Coastal Wetlands for Hurricane Protection. *AMBIO A J. Hum. Environ.* 37, 241–248.
615 [https://doi.org/10.1579/0044-7447\(2008\)37\[241:TVOCWF\]2.0.CO;2](https://doi.org/10.1579/0044-7447(2008)37[241:TVOCWF]2.0.CO;2)
- 616 [13] Maarel, E., 1975. Man-made natural ecosystems in environmental management and planning. *Unifying*
617 *Concepts Ecol.* https://doi.org/https://doi.org/10.1007/978-94-010-1954-5_22
- 618 [14] Escobedo, F.J., Kroeger, T., Wagner, J.E., 2011. Urban forests and pollution mitigation: Analyzing
619 ecosystem services and disservices. *Environ. Pollut.* 159, 2078–2087.
620 <https://doi.org/https://doi.org/10.1016/j.envpol.2011.01.010>
- 621 [15] Everard, M., Jones, L., Watts, B., 2010. Have we neglected the societal importance of sand dunes? An
622 ecosystem services perspective. *Aquat. Conserv. Mar. Freshw. Ecosyst.* 20, 476–487.
623 <https://doi.org/10.1002/aqc.1114>
- 624 [16] Rodríguez, J., García, M., Bombardelli, F., Rhoads, B., Herricks, E., 2019. Naturalization of Urban
625 Streams Using In-Channel Structures. *Build. Partnerships, Proceedings.*
626 [https://doi.org/doi:10.1061/40517\(2000\)341](https://doi.org/doi:10.1061/40517(2000)341)
- 627 [17] Farley, K.A., Jobbágy, E.G., Jackson, R.B., 2005. Effects of afforestation on water yield: a global syn-
628 thesis with implications for policy. *Glob. Chang. Biol.* 11, 1565–1576. [https://doi.org/10.1111/j.1365-](https://doi.org/10.1111/j.1365-2486.2005.01011.x)
629 [2486.2005.01011.x](https://doi.org/10.1111/j.1365-2486.2005.01011.x)
- 630 [18] Fernandes, P.M., Botelho, H.S., 2003. A review of prescribed burning effectiveness in fire hazard re-
631 duction. *Int. J. Wildl. Fire* 12, 117–128.
- 632 [19] Fontana, V., Radtke, A., Walde, J., Tasser, E., Wilhalm, T., Zerbe, S., Tappeiner, U., 2014. What plant
633 traits tell us: Consequences of land-use change of a traditional agro-forest system on biodiversity and
634 ecosystem service provision. *Agric. Ecosyst. Environ.* 186, 44–53.
635 <https://doi.org/https://doi.org/10.1016/j.agee.2014.01.006>
- 636 [20] Francis, C., Jones, A., Crookston, K., Wittler, K., Goodman, S., 1986. Strip cropping corn and grain le-

- 637 gumes: A review. *Am. J. Altern. Agric.* 1, 159–164. <https://doi.org/DOI: 10.1017/S0889189300001235>
- 638 [21] Friess, D., 2016. Ecosystem Services and Disservices of Mangrove Forests: Insights from Historical
639 Colonial Observations, *Forests*. <https://doi.org/10.3390/f7090183>
- 640 [22] Fujita, S., 1997. Measures to promote stormwater infiltration. *Water Sci. Technol.* 36, 289–293.
641 [https://doi.org/https://doi.org/10.1016/S0273-1223\(97\)00584-2](https://doi.org/https://doi.org/10.1016/S0273-1223(97)00584-2)
- 642 [23] Gibson, C.W.D., Hambler, C., Brown, V.K., 1992. Changes in Spider (Araneae) Assemblages in Rela-
643 tion to Succession and Grazing Management. *J. Appl. Ecol.* 29, 132–142.
644 <https://doi.org/10.2307/2404356>
- 645 [24] GreenUP, U., 2018. New Strategy for Re-Naturing Cities through Nature-Based Solutions – NBS Ca-
646 talogue.
- 647 [25] Grimm, N.B., Chapin III, F.S., Bierwagen, B., Gonzalez, P., Groffman, P.M., Luo, Y., Melton, F., Na-
648 delhoffer, K., Pairis, A., Raymond, P.A., Schimel, J., Williamson, C.E., 2013. The impacts of climate
649 change on ecosystem structure and function. *Front. Ecol. Environ.* 11, 474–482.
650 <https://doi.org/10.1890/120282>
- 651 [26] Harris, J., J. Hobbs, R., Higgs, E., Aronson, J., 2006. Ecological Restoration and Global Climate Chan-
652 ge, *Restoration Ecology - RESTOR ECOL.* <https://doi.org/10.1111/j.1526-100X.2006.00136.x>
- 653 [27] Hickey, M.B.C., Doran, B., 2004. A Review of the Efficiency of Buffer Strips for the Maintenance and
654 Enhancement of Riparian Ecosystems. *Water Qual. Res. J.* 39, 311–317.
655 <https://doi.org/10.2166/wqrj.2004.042>
- 656 [28] Hönigová, I., Vačkář, D., Lorencová, E., Melichar, J., Götzl, M., Sonderegger, G., Oušková, V., Hošek,
657 M., Chobot, K., 2012. Survey on grassland ecosystem services Report of the European Topic Centre on
658 Biological Diversity. Prague.
- 659 [29] J. Hobbs, R., 2004. Restoration ecology: the challenge of social values and expectations. *Front. Ecol.*
660 *Environ.* 2, 43–48. [https://doi.org/10.1890/1540-9295\(2004\)002\[0043:RETCOS\]2.0.CO;2](https://doi.org/10.1890/1540-9295(2004)002[0043:RETCOS]2.0.CO;2)
- 661 [30] Kabisch, N., Korn, H., Stadler, J., Bonn, A., 2017. Nature-Based Solutions to Climate Change Adapta-
662 tion in Urban Areas—Linkages Between Science, Policy and Practice BT - Nature-Based Solutions to

- 663 Climate Change Adaptation in Urban Areas: Linkages between Science, Policy and Practice, in:
664 Kabisch, N., Korn, H., Stadler, J., Bonn, A. (Eds.), . Springer International Publishing, Cham, pp. 1–11.
665 https://doi.org/10.1007/978-3-319-56091-5_1
- 666 [31] Kahmen, S., Poschlod, P., Schreiber, K.-F., 2002. Conservation management of calcareous grasslands.
667 Changes in plant species composition and response of functional traits during 25 years. *Biol. Conserv.*
668 104, 319–328. [https://doi.org/10.1016/S0006-3207\(01\)00197-5](https://doi.org/10.1016/S0006-3207(01)00197-5)
- 669 [32] Kniivilä, M., Ovaskainen, V., Saastamoinen, O., Kniivilä, M., 2002. Costs and benefits of forest con-
670 servation: regional and local comparisons in Eastern Finland. *J. For. Econ.* 8, 131–150.
671 <https://doi.org/10.1078/1104-6899-00008>
- 672 [33] Kuo, S., Sainju, U.M., Jellum, E.J., 1997. Winter Cover Crop Effects on Soil Organic Carbon and Car-
673 bohydrate in Soil. *Soil Sci. Soc. Am. J.* 61, 145–152.
674 <https://doi.org/10.2136/sssaj1997.03615995006100010022x>
- 675 [34] Liebman, M., Dyck, E., 1993. Crop Rotation and Intercropping Strategies for Weed Management.
676 *Ecol. Appl.* 3, 92–122. <https://doi.org/10.2307/1941795>
- 677 [35] Locatelli, B., Pavageau, C., Pramova, E., Di Gregorio, M., 2015. Integrating climate change mitigation
678 and adaptation in agriculture and forestry: opportunities and trade-offs. *Wiley Interdiscip. Rev. Clim.*
679 *Chang.* 6, 585–598. <https://doi.org/10.1002/wcc.357>
- 680 [36] Lüderitz, V., Jüpner, R., Müller, S., Feld, C.K., 2004. Renaturalization of streams and rivers — the
681 special importance of integrated ecological methods in measurement of success. An example from Sa-
682 xony-Anhalt (Germany). *Limnologica* 34, 249–263. [https://doi.org/10.1016/S0075-](https://doi.org/10.1016/S0075-9511(04)80049-5)
683 [9511\(04\)80049-5](https://doi.org/10.1016/S0075-9511(04)80049-5)
- 684 [37] Lutz, E., Pagiola, S., Reiche, C., 1994. The costs and benefits of soil conservation: the farmers’
685 viewpoint. *World Bank Res. Obs.* 9, 273–295. <https://doi.org/10.1093/wbro/9.2.273>
- 686 [38] M. Brander, L., J. Wagtendonk, A., S. Hussain, S., McVittie, A., Verburg, P.H., de Groot, R.S., van der
687 Ploeg, S., 2012. Ecosystem service values for mangroves in Southeast Asia: A meta-analysis and value
688 transfer application. *Ecosyst. Serv.* 1, 62–69.
689 <https://doi.org/10.1016/j.ecoser.2012.06.003>

- 690 [39] Mahdi, L., Bell, C.J., Ryan, J., 1998. Establishment and yield of wheat (*Triticum turgidum* L.) after
691 early sowing at various depths in a semi-arid Mediterranean environment. *F. Crop. Res.* 58, 187–196.
692 [https://doi.org/https://doi.org/10.1016/S0378-4290\(98\)00094-X](https://doi.org/https://doi.org/10.1016/S0378-4290(98)00094-X)
- 693 [40] Missimer, T.M., Drewes, J.E., Amy, G., Maliva, R.G., Keller, S., 2012. Restoration of Wadi Aquifers
694 by Artificial Recharge with Treated Waste Water. *Groundwater* 50, 514–527.
695 <https://doi.org/10.1111/j.1745-6584.2012.00941.x>
- 696 [41] Moberg, F., Folke, C., 1999. Ecological goods and services of coral reef ecosystems. *Ecol. Econ.* 29,
697 215–233. [https://doi.org/https://doi.org/10.1016/S0921-8009\(99\)00009-9](https://doi.org/https://doi.org/10.1016/S0921-8009(99)00009-9)
- 698 [42] Moradkhani, H., Baird, R.G., Wherry, S.A., 2010. Assessment of climate change impact on floodplain
699 and hydrologic ecotones. *J. Hydrol.* 395, 264–278.
700 <https://doi.org/https://doi.org/10.1016/j.jhydrol.2010.10.038>
- 701 [43] Mulumba, L.N., Lal, R., 2008. Mulching effects on selected soil physical properties. *Soil Tillage Res.*
702 98, 106–111. <https://doi.org/https://doi.org/10.1016/j.still.2007.10.011>
- 703 [44] Nakano, D., Nakamura, F., 2008. The significance of meandering channel morphology on the diversity
704 and abundance of macroinvertebrates in a lowland river in Japan. *Aquat. Conserv. Mar. Freshw. Eco-*
705 *syst.* 18, 780–798. <https://doi.org/10.1002/aqc.885>
- 706 [45] Nehren, U., Ho Dac, H., Marfai, M.A., Raedig, C., Alfonso de Nehren, S., Junun, S., Castro, C., 2016.
707 Ecosystem Services of Coastal Dune Systems for Hazard Mitigation: Case Studies from Vietnam, Indo-
708 nesia, and Chile. pp. 401–433. https://doi.org/10.1007/978-3-319-43633-3_18
- 709 [46] Nelson, E.J., Kareiva, P., Ruckelshaus, M., Arkema, K., Geller, G., Girvetz, E., Goodrich, D., Matzek,
710 V., Pinsky, M., Reid, W., Saunders, M., Semmens, D., Tallis, H., 2013. Climate change's impact on key
711 ecosystem services and the human well-being they support in the US. *Front. Ecol. Environ.* 11, 483–
712 893. <https://doi.org/10.1890/120312>
- 713 [47] Nesshöver, C., Assmuth, T., Irvine, K.N., Rusch, G.M., Waylen, K.A., Delbaere, B., Haase, D., Jones-
714 Walters, L., Keune, H., Kovacs, E., Krauze, K., Külvik, M., Rey, F., van Dijk, J., Vistad, O.I., Wilkin-
715 son, M.E., Wittmer, H., 2017. The science, policy and practice of nature-based solutions: An inter-
716 disciplinary perspective. *Sci. Total Environ.* 579, 1215–1227.

- 717 <https://doi.org/10.1016/j.scitotenv.2016.11.106>
- 718 [48] NWRM, 2019. Natural Water Retention Measures [WWW Document]. URL <http://www.nwrn.eu> (ac-
719 cessed 2.5.19).
- 720 [49] Osborne, L.L., Kovacic, D.A., 1993. Riparian vegetated buffer strips in water-quality restoration and
721 stream management. *Freshw. Biol.* 29, 243–258. <https://doi.org/10.1111/j.1365-2427.1993.tb00761.x>
- 722 [50] Parton, W.J., Scurlock, J.M.O., Ojima, D.S., Schimel, D.S., Hall, D.O., Members, S.G., 1995. Impact of
723 climate change on grassland production and soil carbon worldwide. *Glob. Chang. Biol.* 1, 13–22.
724 <https://doi.org/10.1111/j.1365-2486.1995.tb00002.x>
- 725 [51] Pedersen, M.L., Kristensen, K.K., Friberg, N., 2014. Re-Meandering of Lowland Streams: Will Dis-
726 obeying the Laws of Geomorphology Have Ecological Consequences? *PLoS One* 9, e108558.
- 727 [52] Piazza, B.P., Allen, Y.C., Martin, R., Bergan, J.F., King, K., Jacob, R., 2014. Floodplain conservation
728 in the Mississippi River Valley: combining spatial analysis, landowner outreach, and market assessment
729 to enhance land protection for the Atchafalaya River Basin, Louisiana, U.S.A. *Restor. Ecol.* 23, 65–74.
730 <https://doi.org/10.1111/rec.12120>
- 731 [53] Piazza, B.P., Banks, P.D., La Peyre, M.K., 2005. The Potential for Created Oyster Shell Reefs as a
732 Sustainable Shoreline Protection Strategy in Louisiana. *Restor. Ecol.* 13, 499–506.
733 <https://doi.org/10.1111/j.1526-100X.2005.00062.x>
- 734 [54] Pitman, A.J., Narisma, G.T., McAneney, J., 2007. The impact of climate change on the risk of forest
735 and grassland fires in Australia. *Clim. Change* 84, 383–401. <https://doi.org/10.1007/s10584-007-9243-6>
- 736 [55] Pollet, J., Omi, P.N., 2002. Effect of thinning and prescribed burning on crown fire severity in pondero-
737 sa pine forests. *Int. J. Wildl. Fire* 11, 1–10.
- 738 [56] Powlson, D.S., Stirling, C.M., Jat, M.L., Gerard, B.G., Palm, C.A., Sanchez, P.A., Cassman, K.G.,
739 2014. Limited potential of no-till agriculture for climate change mitigation. *Nat. Clim. Chang.* 4, 678.
- 740 [57] Scyphers, S.B., Powers, S.P., Heck Jr, K.L., Byron, D., 2011. Oyster Reefs as Natural Breakwaters Mi-
741 tigate Shoreline Loss and Facilitate Fisheries. *PLoS One* 6, e22396.

- 742 [58] Seabloom, E.W., Ruggiero, P., Hacker, S.D., Mull, J., Zarnetske, P., 2013. Invasive grasses, climate
743 change, and exposure to storm-wave overtopping in coastal dune ecosystems. *Glob. Chang. Biol.* 19,
744 824–832. <https://doi.org/10.1111/gcb.12078>
- 745 [59] Seddon, N., Turner, B., Berry, P., Chausson, A., Girardin, C.A.J., 2019. Grounding nature-based clima-
746 te solutions in sound biodiversity science. *Nat. Clim. Chang.* 9, 84–87. [https://doi.org/10.1038/s41558-](https://doi.org/10.1038/s41558-019-0405-0)
747 [019-0405-0](https://doi.org/10.1038/s41558-019-0405-0)
- 748 [60] SER, P., 2002. SER Primer on Ecological Restoration [WWW Document]. 2002. URL www.ser.org
- 749 [61] Shixiong, C., 2008. Why large-scale afforestation efforts in China have failed to solve the desertification
750 problem.
- 751 [62] Stanchi, S., Freppaz, M., Agnelli, A., Reinsch, T., Zanini, E., 2012. Properties, best management prac-
752 tices and conservation of terraced soils in Southern Europe (from Mediterranean areas to the Alps): A
753 review. *Quat. Int.* 265, 90–100. [https://doi.org/https://doi.org/10.1016/j.quaint.2011.09.015](https://doi.org/10.1016/j.quaint.2011.09.015)
- 754 [63] Stoffels, R.J., Clarke, K.R., Rehwinkel, R.A., McCarthy, B.J., 2013. Response of a floodplain fish
755 community to river-floodplain connectivity: natural versus managed reconnection. *Can. J. Fish. Aquat.*
756 *Sci.* 71, 236–245. <https://doi.org/10.1139/cjfas-2013-0042>
- 757 [64] Temmerman, S., Meire, P., Bouma, T.J., Herman, P.M.J., Ysebaert, T., De Vriend, H.J., 2013. Ecosys-
758 tem-based coastal defence in the face of global change. *Nature* 504, 79.
- 759 [65] Thinknature, n.d. Thinknature-platform for NBS [WWW Document]. URL [https://www.think-](https://www.think-nature.eu/)
760 [nature.eu/](https://www.think-nature.eu/) (accessed 2.1.19).
- 761 [66] Uusitalo, P., Lankoski, J., Ollikainen, M., 2006. No-till technology: benefits to farmers and the en-
762 vironment? Theoretical analysis and application to Finnish agriculture. *Eur. Rev. Agric. Econ.* 33, 193–
763 221. <https://doi.org/10.1093/erae/jbl003>
- 764 [67] Vymazal, J., 2007. Removal of nutrients in various types of constructed wetlands. *Sci. Total Environ.*
765 380, 48–65. [https://doi.org/https://doi.org/10.1016/j.scitotenv.2006.09.014](https://doi.org/10.1016/j.scitotenv.2006.09.014)
- 766 [68] Walles, B., Troost, K., van den Ende, D., Nieuwhof, S., Smaal, A.C., Ysebaert, T., 2016. From artificial
767 structures to self-sustaining oyster reefs. *J. Sea Res.* 108, 1–9.

- 768 <https://doi.org/https://doi.org/10.1016/j.seares.2015.11.007>
- 769 [69] Woodward, R.T., Wui, Y.-S., 2001. The economic value of wetland services: a meta-analysis. *Ecol.*
770 *Econ.* 37, 257.
- 771 [70] Zhu, H., Fu, B., Wang, S., Zhu, L., Zhang, L., Jiao, L., Wang, C., 2015. Reducing soil erosion by im-
772 proving community functional diversity in semi-arid grasslands. *J. Appl. Ecol.* 52, 1063–1072.
773 <https://doi.org/10.1111/1365-2664.12442>
- 774 [71] Zhu, K., Woodall, C.W., Clark, J.S., 2011. Failure to migrate: lack of tree range expansion in response
775 to climate change. *Glob. Chang. Biol.* 18, 1042–1052. [https://doi.org/10.1111/j.1365-](https://doi.org/10.1111/j.1365-2486.2011.02571.x)
776 [2486.2011.02571.x](https://doi.org/10.1111/j.1365-2486.2011.02571.x)
- 777 [72] Zougmore, R., Kambou, F.N., Ouattara, K., Guillobez, S., 2000. Sorghum-cowpea Intercropping: An
778 Effective Technique Against Runoff and Soil Erosion in the Sahel (Saria, Burkina Faso). *Arid Soil Res.*
779 *Rehabil.* 14, 329–342. <https://doi.org/10.1080/08903060050136441>
- 780
- 781
- 782
- 783

