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**Occurrence and spatial distribution of organophosphate ester
flame retardants and plasticizers in 40 rivers draining into the
Bohai Sea, north China**

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1 Occurrence and screening-level risk assessment of
2 organophosphate ester flame retardants and plasticizers in
3 40 rivers around the Bohai Sea, north China

4 *Runmei Wang,^{†,§} Jianhui Tang,^{*,†} Zhiyong Xie,^{‡***} Wenying Mi,[‡] Yingjun Chen,[†] Hendrik*
5 *Wolschke[‡], Chongguo Tian,[†] Xiaohui Pan,[†] Yongming Luo,[†] Ralf Ebinghaus,[‡]*

6 [†]Key Laboratory of Coastal Environmental Processes and Ecological Remediation, Yantai
7 Institute of Coastal Zone Research, CAS, Yantai, 264003, China

8 [‡]Helmholtz-Zentrum Geesthacht, Centre for Materials and Coastal Research, Institute of Coastal
9 Research, Max-Planck-Strasse 1, Geesthacht, 21502, Germany

10 [§]University of Chinese Academy of Sciences, Beijing 100039, China

11 **ABSTRACT:** As an alternative of polybrominated diphenyl ethers (PBDEs), environmental
12 level and risk assessment about organophosphate esters (OPEs) are required as the worldwide
13 demand for OPEs has been increasing every year. This study focused on the spatial distribution
14 and screening-level risk assessment of organophosphate ester flame retardants and plasticizers in
15 40 major rivers entering into the Bohai Sea, north China. Water samples were filtrated by glass
16 fiber filters, then extracted by liquid-liquid extraction, and analyzed by gas chromatography-mass
17 spectrometry (GC-MS) for 15 different OPE congeners and one synthetic intermediate
18 triphenylphosphine oxide (TPPO). The sum of the OPE (Σ OPE) concentrations in 40 rivers

19 ranged from 10 to 1566 ng L⁻¹, with the mean ΣOPE concentration of 344 ng L⁻¹. Tri(1-chloro-2-
20 propyl) phosphate (TCPP) (range: 5-921ng L⁻¹, with a mean concentration of 186 ng L⁻¹) and
21 tri(2-chloroethyl) phosphate (TCEP) (range:1 – 418 ng L⁻¹, mean:88 ng L⁻¹) were the most
22 abundant OPEs in river waters of North China. It is noted that the concentration levels of TPPO
23 (range: <LOD – 5852 ng L⁻¹, mean: 224 ng L⁻¹) were much higher than those reported in the
24 Elbe, Germany. The riverine inputs of target organophosphorus compounds were estimated to be
25 8.0 t yr⁻¹ for Liaodong Bay, 3.6 t yr⁻¹ for Bohai Bay and 1.7 t yr⁻¹ for Laizhou Bay, respectively.
26 In addition, a screening-level risk assessment was conducted with RQ ranging from 0.0004 to
27 0.1129, which indicated no significant risk presented in all the 40 rivers. However, for these
28 potentially persistent and bioaccumulative compounds, a refined and long-term risk assessment
29 needs to be carried on.

30 **Brief:** Organophosphate ester flame retardants and plasticizers are ubiquitous in the aquatic
31 environment and subject to emerging contaminants discharged into the coastal environment.

32 INTRODUCTION

33 Organophosphate esters (OPEs) have been used as flame retardants and plasticizers for decades
34 due to their excellent physicochemical properties. Since brominated flame retardants (BFRs)
35 have been gradually restricted worldwide due to the environmental concerns, OPEs, as an
36 alternative of BFRs, have been increasingly used in many industrial applications and household
37 products. Halogenated (chlorinated) OPEs are predominantly used as flame retardants in
38 furniture, textiles, mattresses, electronics (e.g., televisions, cell phones), and even children
39 products such as strollers, sleepwear and baby clothing.¹⁻⁵ Non-halogenated OPEs are mostly
40 applied as plasticizers, antifoaming agents and additives.^{1,2} Another organophosphorus

41 compound triphenylphosphine oxide (TPPO) is extensively employed as synthetic intermediate
42 in pharmaceutical products and as ligand for many transitional metals.⁶

43 Organophosphorus flame retardants are listed in the High Production Volume Chemicals
44 (HPVC) programme.⁷ The global consumption of OPEs amounted to 500 000 t in 2011 and
45 expected to reach 680.000 t in 2015.⁸ In China, the production of PFRs (phosphorus flame
46 retardants) was estimated to be 100 000 t in 2011 and demands for PFRs was expected to
47 increase every year by 15%.⁸ Since OPEs are not chemically bonded to polymeric materials and
48 typically water soluble, they can easily leach out of the material into the environment via
49 volatilization, abrasion, and dissolution.⁹ Large production and consumption of OPEs resulted in
50 high detection frequency in both domestic circumstance (air and dust) and nature environment
51 (water, air and sediment) over the past decade.^{9,10}

52 Halogenated OPEs have been proved to be persistent in the environment and resistant to
53 hydrolysis in neutral pH.¹⁰ Risk assessment with respect to human health for OPEs has not yet
54 been completed.^{10,11} Nevertheless, the draft on the risk assessment of tri(2-chloroethyl)phosphate
55 (TCEP) recognized it's carcinogenicity, high toxicity and environmental persistence.¹²
56 Furthermore, tri (dichloropropyl) phosphate (TDCPP) had been proved to be carcinogenic, and
57 tri(1-chloro-2-propyl) phosphate (TCPP) and tributoxyethyl phosphate (TBEP) were also
58 suspected carcinogens.^{13,14} New York prohibited use of TCEP in children care products under
59 the age of 3 from December 1, 2013.¹⁵ After New York city, Washington will prohibit the use of
60 TCEP and TDCPP in children products and home furniture in July 1, 2014.¹⁶

61 Considering the toxic effects of these compounds, together with the limited reports on the
62 occurrence of OPEs in the environment, it is worthy to collect more information on these
63 contaminants to fill our knowledge gaps of the fate of OPEs in the environment.

64 Bohai Sea is an enclosed inner sea of China, with huge amount of domestic sewage and
65 industrial waste water pouring into it every day. Rivers are the major source/pathway responsible
66 for the pollutants into the sea. According to a recent report about the water quality in the rivers
67 emptying into the China coastal seas, all the 11 routine monitored rivers emptying into the Bohai
68 Sea were equal to or lower than Level IV National Environmental Quality Standards for Surface
69 Water, 5 of 11 rivers had even exceed the Level V standards.¹⁷

70 To investigate the distribution characteristics and fluxes of emerging organic compounds in the
71 rivers around the Bohai Sea, 40 major rivers were selected. The rivers were chosen based on their
72 length, water loading volume, pollution status according to previous studies and the total water
73 volume of these 40 rivers accounted for about 96 % of these rivers emptying into the Bohai Sea
74 and northern Yellow Sea, for detailed information see [Table S1](#).

75 In the river basin areas, it has been subject to heavy anthropogenic influences owing to the
76 high speed development of the agriculture, industry and economy during the past ca. 50 years.¹⁸
77 There is one of the most important heavy industrial complexes in Northeast Asia, including
78 chemical (flame retardants production), petrochemical, pharmaceutical, steel-iron, and machinery
79 industries.¹⁸ Nearly 100 billion tons of river water carried kinds of organic pollutants including
80 OPEs flowing into the Bohai Sea in a year. Moreover, phosphorus is a key nutrient for marine
81 phytoplankton and microbial communities, OPEs as a kind of organic phosphorus may play an
82 important role in biogeochemical cycle of phosphorus.¹⁹ In the Bohai Sea, the limiting nutrients
83 of phytoplankton was changed from nitrogen in the early 1980s to nitrogen-phosphorus in the
84 late 1980s, and then to phosphorus after the 1990s.²⁰ Therefore, it is more worthy to estimate the
85 riverine input of organic phosphorus such as OPEs.

86 In the present study, OPEs were analyzed to determine their occurrence in aqueous phase and
87 their profiles in different rivers. In addition, the riverine input of OPEs into the Bohai Sea was

88 estimated and a risk assessment was conducted to evaluate the influence of OPEs. To the best of
89 our knowledge, this is the first study of organophosphate esters (OPEs) in aqueous environment
90 around Bohai Sea and the first report of environmental distribution of TPPO in China.

91 **EXPERIMENTAL SECTION**

92 **Study area and sampling.** Water samples were collected in 40 major rivers around the Bohai
93 Sea and the north Yellow Sea in August 2013. The locations of sampling sites were shown in
94 [Figure 1](#). The total watershed area is up to 1 412 581 km², nearly 15% of China ([Table S1](#)).²¹
95 The water sampling procedure was conforming to the Chinese National Standard -Technical
96 Specifications Requirements for Monitoring of Surface Water and Waste Water (HJ/T91 2012)
97 and changed with small modifications. In brief, sampling sites were chosen as close as the river
98 mouth but avoided to the influence of salt water. In most of the rivers in this study, there are
99 floodgates/dams near the river mouth to prevent the sea water intrusion during high tide period
100 and also to keep the fresh water level high. Water samples were collected across the
101 floodgates/dams and in some cases across the bridges. In each river, one transection with 3 to 5
102 points according to the river width was set, and a 10 L stainless steel bucket was used to collect
103 the surface water. Three to five 10-L water samples were mixed together in one 60 L stainless
104 steel bucket and one liter of mixed water was kept frozen in PET bottles prior to extraction. All
105 the sampling buckets and PET bottles were rinsed with acetone in a clean-lab, and then river
106 water three times on the spot.

107 **Analysis.** The water samples were filtrated by glass fiber filters (GFF, Φ: 47 mm, pore size:
108 0.7 μm). The filtrate was added with an aliquot of surrogate standard solution (D₂₇ TnBP and D₁₅
109 TPhP) and then extracted with 50 ml dichloromethane for 3 times. The extracts were combined
110 and then concentrated to 150 μL for GC-MS analysis. 500 pg ¹³C₆-PCB 208 was added into each

111 sample as injection standard. The samples were analyzed by GC-MS (Agilent 6890 gas
112 chromatograph coupled to an Agilent 5975 mass spectrometer) equipped with a PTV
113 (programmed temperature vaporizer) injector. The GC separation was performed with a HP-5MS
114 column (length: 30 m, ID: 0.25 mm; film: 0.25 μm , J&W Scientific). The PTV (2 μL injection
115 volume with a pulse pressure of 20 psi for 2 min and inlet temperature of 280 $^{\circ}\text{C}$) was operated
116 in PTV pulse splitless mode 50 $^{\circ}\text{C}$ (2 min) \rightarrow 5 $^{\circ}\text{C}/\text{min}$ \rightarrow 170 $^{\circ}\text{C}$ (5 min) \rightarrow 10 $^{\circ}\text{C}/\text{min}$ \rightarrow
117 230 $^{\circ}\text{C}$ (5 min) \rightarrow 5 $^{\circ}\text{C}/\text{min}$ \rightarrow 270 $^{\circ}\text{C}$ \rightarrow 30 $^{\circ}\text{C}/\text{min}$ \rightarrow 300 $^{\circ}\text{C}$ (10 min). The quadrupole was
118 maintained at 150 $^{\circ}\text{C}$ and the ion source of the mass spectrometer was operated at 230 $^{\circ}\text{C}$ at 70
119 eV electron impact (EI). Selected masses of fragment ions for quantification and quantitation
120 were listed in Table S1. The response factors were derived from the calibration curves (8-points)
121 made for response ratio between target compounds and surrogate standards (0-5000 ng mL^{-1}).

122 This study covered 16 organophosphorus compounds: three halogenated (chlorinated) alkyl
123 phosphates, including tris(1-chloro-2-propyl) phosphate (TCPP), tris(2-chloroethyl) phosphate
124 (TCEP), and tris(dichloroisopropyl) phosphate (TDCPP); ten non-halogenated alkyl phosphates,
125 including trimethyl phosphate (TMP), triethyl phosphate (TEP), tri-*iso*-propyl phosphate (TiPrP),
126 tri-*n*-propyl phosphate (TPrP), tri-*iso*-butyl phosphate (TiBP), tri-*n*-butyl phosphate (TBP),
127 tripentyl phosphate (TPeP), trihexyl phosphate (THP), tris(2-butoxyethyl) phosphate (TBEP),
128 and tris(2-ethylhexyl) phosphate (TEHP); two aryl phosphates, including tricresyl phosphate
129 (TCP), and triphenyl phosphate (TPP) and also the synthetic intermediate triphenylphosphine
130 oxide (TPPO). Their acronym, chemical structures, applications, toxicity and CAS No. were
131 given in [Table 1](#).

132 **Quality Assurance/Quality Control (QA/QC).** The recoveries of spiked experiments were
133 from $67 \pm 2\%$ (TPPO) to $95 \pm 8\%$ (TCEP) for 15 target organophosphorus compounds with a
134 mean recovery of 81.4% ($n=5$), except from DMP ($36 \pm 2\%$). A method blank with each sample

135 batch (six samples) was included. The mean blanks were between 7 ± 3 (TPPO) to 235 ± 102
136 (TCPP) pg/L. The instrumental limit of detection (LOD) was defined as three times the signal-to-
137 noise ($S/N=3$) and ranged from 0.1 pg (TnBP) to 6 pg (TBEP). The method detection limit (MDL)
138 was defined as the mean field blank concentrations plus three times the standard deviation (3σ) of
139 the field blanks and ranged from 0.018 (TPPO) to 0.55 (TCPP) pg/L (Table S2). All
140 concentrations of organophosphorus compounds in this paper were not corrected for recoveries.

141 RESULTS AND DISCUSSION

142 **Concentration and distribution characteristics.** OPEs were detected in all samples with water
143 phase concentrations decreasing as follows: TCPP ($4.59\sim 921$ ng L⁻¹, mean: 187.20 ng L⁻¹), TCEP
144 ($1.29\sim 268$ ng L⁻¹, mean: 81 ng L⁻¹), TEP (<LOD~350 ng L⁻¹, mean: 43.6 ng L⁻¹), TiBP
145 (<LOD~218 ng L⁻¹, mean: 13 ng L⁻¹). TPPO was up to 5850 ng L⁻¹ in concentration. The detailed
146 concentration data was listed in Table 2. Fig 2 presents the concentrations of TCPP, TCEP and
147 TPPO in sampling rivers from northeast (Yalu River) to south (Jiahe River) around Bohai Sea.
148 The distribution patterns of TCPP and TCEP are similar, which may indicate they came from the
149 same source. What was worth noting, TPPO was detected with concentration <LOD to 370 ng L⁻¹
150 ¹ in the samples except Jiehe River and Yalu River. It was up to 1280 ng L⁻¹ and 5850 ng L⁻¹ in
151 Jiehe River and Yalu River, respectively, far higher than the rest of the rivers. In terms of rivers,
152 Yalu River, Jiehe River, Jiahe River, Xiaoling River and Liugu River were the most
153 contaminated by OPEs and TPPO, with a total concentration of 6079 ng L⁻¹, 2300 ng L⁻¹, 1590
154 ng L⁻¹, 1350 ng L⁻¹, and 1100 ng L⁻¹, respectively.

155 TCPP and TCEP were the most abundant OPEs in most of the rivers, while the sum of their
156 concentrations ([TCPP] + [TCEP]) ranged from 41 to 98% (average 80%) of the Σ_{15} OPEs
157 compositions (Fig 3). This behavior is in agreement with studies in Pearl River.⁹ It may be

158 attribute to that TCPP and TCEP were two of the most common products of halogenated
159 phosphate esters and that they appeared to be the most recalcitrant in water.¹⁰ However, TCEP
160 was not amid the most abundant OPEs in German rivers, for instance, Elbe, Rhine and four river
161 systems in Hessen.^{22, 23} The difference could be explained by the industrial replacement of TCEP
162 by TCPP in Europe in the 1990s.²⁴ In this study, there was no reflection of the shift in usage from
163 TCEP to TCPP in China. TEP, TiBP, TDCPP, TBP, TPP and TBEP presented intermediate
164 concentration, ranging from <LOD to 350 ng L⁻¹. It may reflect the widespread distribution of
165 this family of contaminants in rivers influenced by anthropogenic pressures, as reported for other
166 sites previously. TMP, TiPrP, TPrP, TPeP, THP, TEHP and TCP were detected at the lowest
167 concentration or not detected out in most of the rivers. The different pattern of OPEs may
168 attribute to their difference in physicochemical properties and consequent difference in
169 accumulation features and degradability as well as difference in production and employment.

170 In addition, 37 of sampling rivers flowing into three bays of the Bohai Sea, namely, Liaodong
171 Bay, Bohai Bay and Laizhou Bay. It could be shown that the average concentrations of OPEs in
172 rivers flowing into Laizhou Bay (681 ng L⁻¹) were 1.2 times higher than in rivers flowing into
173 Liaodong Bay (545 ng L⁻¹) and more than 3.8 times higher than in rivers flowing into Bohai Bay
174 (180 ng L⁻¹). It was suggested that rivers flowing through the OPEs production sources
175 contributed to the high values.

176 There are very limited reports about TPPO in river water. Still, concentration of TPPO in this
177 study (0.70-5850 ng L⁻¹) was far higher than in Elbe (Germany, 10-40 ng L⁻¹) and in three
178 Volcanic Lakes (Italy, 2 ± 1 ng L⁻¹).^{1, 25} Although the concentration of TPPO was much lower in
179 contrast to this study it was assessed as one of the major organophosphorus compound. Since the
180 study area was one of the most important heavy industrial complexes in China, including

181 chemical, petrochemical, pharmaceutical, steel-iron, and machinery industries,¹⁸ it may exert
182 severe pressure on the environment.

183 To our knowledge, limited reports about the occurrence of OPEs in river water were available
184 throughout the world. Concentrations of selected OPEs and TPPO in others studies in river water
185 were listed in Table 3. In contrast to those published previously, organophosphorus compounds
186 contamination in China maintained high level but not higher than UK. What's more, priority
187 attention should be given to TPPO contamination.

188 **Riverine input into the Bohai Sea.** According to the samples concentrations and the annual
189 runoff, the total riverine input of OPEs was estimated to be 18 t yr⁻¹ and the total riverine input of
190 TPPO was estimated to be 113 t yr⁻¹(Table S3). It's generally considered that riverine runoff
191 plays an important role for the transportation of anthropogenic pollutants from terrestrial source
192 to the ocean.²⁶ Liaodong Bay received relatively high riverine input of OPEs and TPPO with 8.0 t
193 yr⁻¹ in comparison to the Bohai Bay (3.6 t yr⁻¹) and Laizhou Bay (1.7 t yr⁻¹).

194 As we all know, river discharges are an important nutrient sources including phosphate for
195 Bohai Sea. Especially for the phosphorus limitation trend of Bohai Sea, not only inorganic
196 phosphate from riverine input is important but also the organic phosphate, for instance,
197 organophosphorus compound, may make a difference.²⁰

198 **Risk assessment.** As discussed earlier, China is facing severe organophosphorus compounds
199 contamination. The risk assessment for river water organisms was estimated according to the
200 water concentration of the most abundant and intermediate concentration organophosphorus
201 compounds (TCPP, TCEP, TPPO, TEP, TiBP, TDCPP, TBP, TPP, and TBEP) and L(E)C₅₀ of
202 *Daphnia magna* and fish (only for TPPO), following the recommendation of the Technical
203 Guidance Document on Risk Assessment.²⁷ Risk quotients were calculated according to:

$$RQ = \frac{MEC}{PNEC} = \frac{MEC}{EC_{50}/f}$$

204
205 Where MEC means environmental concentration, PNEC is the predicted no effect concentration,
206 and f is the security factor or uncertainty factor. A value of 1000 is adopted in this work.²⁷ Data
207 was interpreted by the maximum probable risk for ecological effects from river water was
208 followed as recommended by.²⁸ It indicates no significant risk for $RQ < 1.0$, indicates a small
209 potential for adverse effects for $1.0 \leq RQ < 10$, and indicates significant potential for adverse
210 effects for $10 \leq RQ < 100$. In other words, the exposure point concentration is equal to or one
211 hundred times greater than the risk screening benchmark. It is indicated that potential adverse
212 effects should be expected when RQ equal of great than 100.

213 The toxicological relevant concentrations (LC50 or EC50) used for RQ calculations are
214 summarized in [Table S4](#), and the RQs results calculated for each compound in the rivers are
215 listed [Table S5](#). RQs for TCPP, TCEP, TEP, TiBP, TDCPP, TBP, TPP, and TBEP varied from 0
216 to 0.02 for daphnia magna. RQs for TPPO varied from 0.00001 to 0.11 for fish. As a
217 consequence, no significant risk ($RQ < 1$) was estimated along all the 40 rivers for daphnia
218 magna (TPPO for fish).

219 Although no significant risk was observed to the aquatic organisms considering short-term
220 exposure, organophosphorus compounds are potentially persistent in water. Concentration of
221 TCEP is still relatively high in Europe ([Table 3](#)) even though it had been replaced by TCPP 20
222 years ago. Moreover, it had been proved that organophosphorus flame retardants can be
223 accumulated in biota. Therefore, more detection and risk assessment of organophosphorus
224 compounds in aquatic ecosystem for either short-term or long-term should be carried on.

225 **Environmental implication.** This study was carried out screening-level occurrence and risk
226 assessment for OPEs in rivers flowing into the Bohai Sea, which located in one of the most

227 important heavy industrial complexes in Northeast Asia with heavy anthropogenic influences
228 owing to the high speed development of the agriculture, industry and economy during the past 50
229 years. Chlorinated OPE, e.g. TCPP and TCEP were the most abundant OPEs in river waters of
230 North China, indicating the influence of large production and consumption of chlorinated OPE in
231 the regions around Bohai Sea. Priority should be given to TPPO due to its relatively high
232 concentration in river waters and the report that it is harmful to aquatic organisms, may cause
233 long-term adverse effects in the aquatic environment. Though no significant risk was estimated
234 for OPEs along all the 40 rivers for daphnia magna (TPPO for fish), considering the large
235 production and consumption of OPEs in China and the fact that New York and Washington have
236 already restricted the usage of TCEP and TDCPP, there is a need to conduct more studies on
237 occurrence and risk assessment of organophosphate ester (OPE) flame retardants and plasticizers
238 in the aquatic ecosystem and coastal environment.

239 ASSOCIATED CONTENT

240 **Supporting information**

241 Additional figures and tables are given in the Supporting Information. This material is available
242 free of charge via the Internet at <http://pubs.acs.org>.

243 AUTHOR INFORMATION

244 **Corresponding Author**

245 * phone: +86-535-2109151; fax: +86-535-2109000; e-mail: jhtang@yic.ac.cn

246 ** phone: +49-4152-872330; Fax: +49-4152-872332; e-mail: zhiyong.xie@hzg.de

247 **Notes**

248 The authors declare no competing financial interest.

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352

353 **Figure caption**

354 **Figure 1:** Sampling sites in the 40 rivers around the Bohai Sea

355 **Figure 2:** Occurrence of TCPP, TCEP and TPPO in 40 sampling rivers in ng L^{-1} around the
356 Bohai sea from northeast (Yalu River) to south (Jiahe River)

357 **Figure 3:** Percentage composition of 15 OPEs in the river water

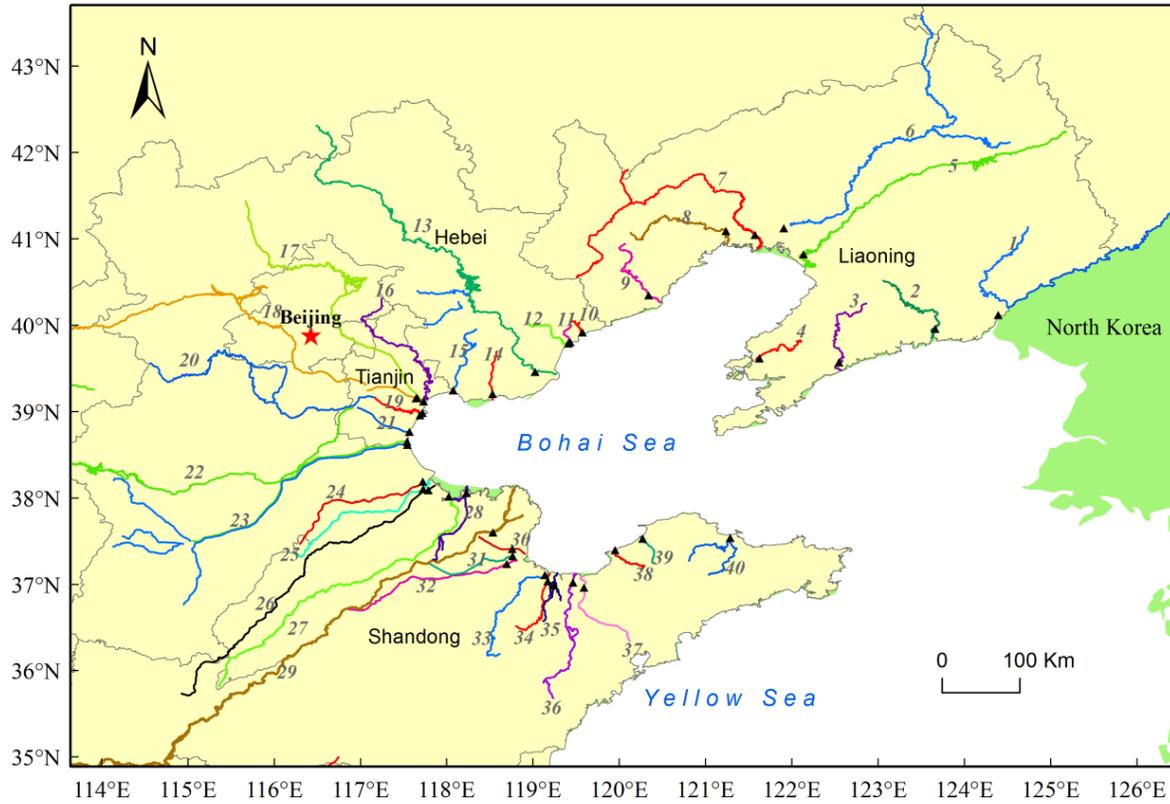
358 **Table title**

359 **Table 1:** The chemical structures, applications, toxicity, CAS No. and physicochemical.
360 properties ($\log K_{OW}$: n-octanol-water partition coefficient, WS: water solubility, VP: vapor
361 pressure) of the OPEs and TPPO in this study

362 **Table 2:** 16 target organophosphorus compounds concentrations in 40 rivers around Bohai Sea,
363 north China

364 **Table 3.** Concentrations of selected OPEs and TPPO in different studies in river water in the
365 world (ng L^{-1})

Figure 1:



No.	river	No.	river
1	Yalu River	21	Dagupaiwu River
2	Dayang River	22	Ziya River
3	Biliu River	23	Beipai River
4	Fuzhou River	24	Xuanhui River
5	Daliao River	25	Zhangweixin River
6	Liaoh River	26	Majia River
7	Daling River	27	Tuhai River
8	Xiaoling River	28	Chaohe River
9	Liugu River	29	Yellow River
10	Tanghe River	30	Guangli River
11	Daihe River	31	Zimai River
12	Yanghe River	32	Xiaoqing River
13	Luanhe River	33	Mihe River
14	Xiaoqinglong River	34	Bailang River
15	Douhe River	35	Yuhe River
16	Jiyun River	36	Weihe River
17	Chaobai River	37	Jiaolai River
18	Yongdingxin River	38	Wanghe River
19	Haihe River	39	Jiehe River
20	Duliujian River	40	Jiahe River

Figure 2:

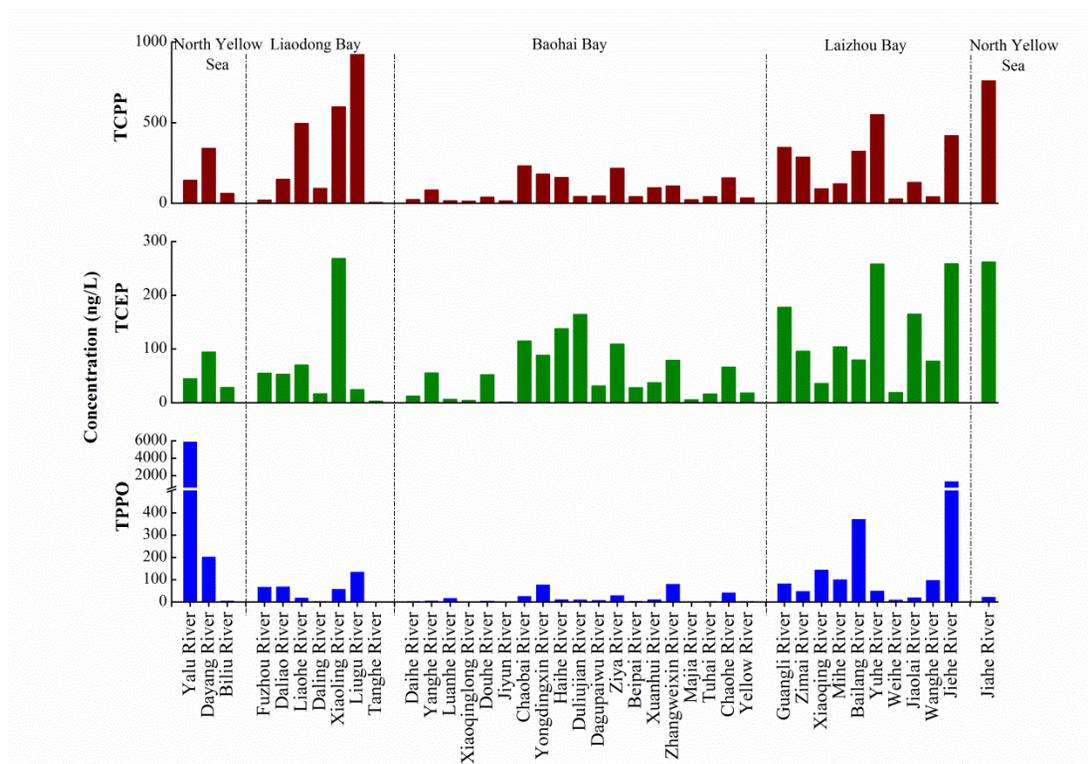


Figure 3:

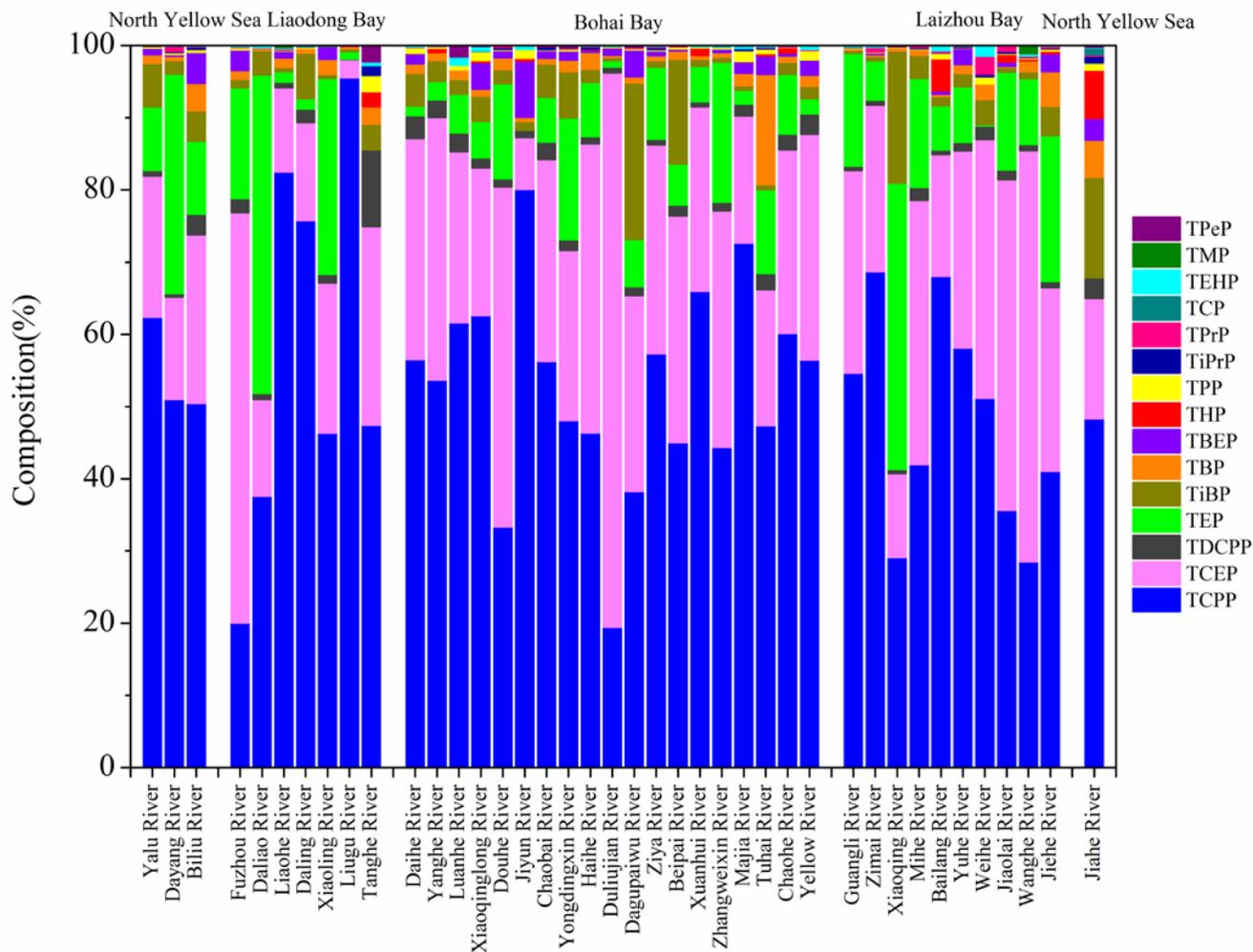
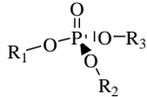


Table 1. The chemical structures, applications, toxicity, CAS No. and physicochemical. properties (log K_{OW} : n-octanol-water partition coefficient, W_S : water solubility, V_P : vapor pressure) of the OPEs and TPPO in this study

Acronym	Chemical structure 	Applications 1, 6, 29	Toxicity 13, 14, 30, 4, 31	phys.-chem. properties ¹	CAS no. ¹
TCPP	$R_1=R_2=R_3=$ 	flame retardant, plasticizer	suspected carcinogenicity ; bioaccumulation; suspected toxicity;	log $K_{OW} = 2.59$ $W_S = 1.2 \text{ g L}^{-1}$ $V_P = 2.7 \cdot 10^{-3} \text{ Pa}$	13674-84-5
TCEP	$R_1=R_2=R_3=$ 	flame retardant, plasticizer, lacquer/paint/glue, industrial processes	carcinogenicity ; neurotoxicity; suspected toxicity;	log $K_{OW} = 1.44$ $W_S = 7.0 \text{ g L}^{-1}$ $V_P = 8.2 \text{ Pa}$	115-96-8
TDCPP	$R_1=R_2=R_3=$ 	flame retardant, plasticizer, lacquer/paint/glue	carcinogenicity; neurotoxicity; higher acute toxicity than TCEP and TCPP	log $K_{OW} = 3.65$ $W_S = 7.0 \cdot 10^{-3} \text{ g L}^{-1}$ $V_P = 9.8 \cdot 10^{-6} \text{ Pa}$	13674-87-8
TMP	$R_1=R_2=R_3=$ $-\text{CH}_3$	industrial processes(pharmaceutical, pesticide solvent and extractant)	n.a.	log $K_{OW} = -0.65$ $W_S = 500 \text{ g L}^{-1}$ $V_P = 113 \text{ Pa}$	512-56-1
TEP	$R_1=R_2=R_3=$ 	flame retardant, plasticizer, industrial processes(raw materials of pesticides)	n.a.	log $K_{OW} = 0.8$ $W_S = 500 \text{ g L}^{-1}$ $V_P = 52.4 \text{ Pa}$	78-40-0
TiPrP	$R_1=R_2=R_3=$ 	n.a.	n.a.	log $K_{OW} = 2.12$ $W_S = 0.50 \text{ g L}^{-1}$ $V_P = 18.4 \text{ Pa}$	513-02-0
TPrP	$R_1=R_2=R_3=$ 	n.a.	n.a.	log $K_{OW} = 1.87$ $W_S = 6.5 \text{ g L}^{-1}$ $V_P = 0.58 \text{ Pa}$	513-08-6
TiBP	$R_1=R_2=R_3=$ 	lacquer/paint/glue, anti-foaming agent, industrial processes	n.a.	log $K_{OW} = 3.6$ $W_S = 1.6 \cdot 10^{-2} \text{ g L}^{-1}$ $V_P = 1.71 \text{ Pa}$	126-71-6
TBP	$R_1=R_2=R_3=$ 	plasticizer, hydraulic fluid, floor covering, lacquer/paint/glue, anti-foaming agent,	suspected neurotoxicity	log $K_{OW} = 4.0$ $W_S = 0.28 \text{ g L}^{-1}$	126-73-8

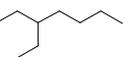
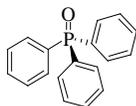
		industrial processes		$V_p = 0.15 \text{ Pa}$	
TPeP	$R_1=R_2=R_3=$ 	n.a.	n.a.	$\log K_{OW} = 5.29$ $W_S = 3.3 \cdot 10^{-4} \text{ g L}^{-1}$ $V_p = 2.2 \cdot 10^{-3} \text{ Pa}$	2528-38-3
THP	$R_1=R_2=R_3=$ 	n.a.	n.a.	n.a.	2528-39-4
TBEP	$R_1=R_2=R_3=$ 	flame retardant, plasticizer, fungus resistance, lacquer/paint/glue, anti-foaming agent	suspected carcinogenicity	$\log K_{OW} = 3.75$ $W_S = 1.1 \text{ g L}^{-1}$ $V_p = 3.3 \cdot 10^{-6} \text{ Pa}$	78-51-3
TEHP	$R_1=R_2=R_3=$ 	flame retardant, plasticizer, fungus resistance	n.a.	$\log K_{OW} = 9.49$ $W_S = 6.0 \cdot 10^{-4} \text{ g L}^{-1}$ $V_p = 1.1 \cdot 10^{-5} \text{ Pa}$	78-42-2
TCP	$R_1=R_2=R_3=$ 	flame retardant, hydraulic fluid, lacquer/paint/glue, industrial processes	n.a.	$\log K_{OW} = 5.11$ $W_S = 3.6 \cdot 10^{-4} \text{ g L}^{-1}$ $V_p = 8.0 \cdot 10^{-5} \text{ Pa}$	1330-78-5
TPP (TPhP)	$R_1=R_2=R_3=$ 	flame retardant, plasticizer, hydraulic fluid, lacquer/paint/glue	suspected neurotoxicity; high toxicity; bioaccumulation	$\log K_{OW} = 4.59$ $W_S = 1.9 \cdot 10^{-3} \text{ g L}^{-1}$ $V_p = 8.4 \cdot 10^{-4} \text{ Pa}$	115-86-6
TPPO		flame-retardant; ligand for many metals; solvent extraction; synthetic intermediate Crystallization Aid	Harmful to aquatic organisms, may cause long-term adverse effects in the aquatic environment	$\log K_{OW} = 2.83$ $W_S = 6.3 \cdot 10^{-2} \text{ g L}^{-1}$ $V_p = 3.47 \cdot 10^{-7} \text{ Pa}$	791-28-6

Table 2. 16 target organophosphorus compounds concentrations in 40 rivers around Bohai Sea, north China

Compound	Min (ng L⁻¹)	Max (ng L⁻¹)	Mean (ng L⁻¹)	Median (ng L⁻¹)	Detection rate (%)
TCPP	4.59	921	186	101	100
TCEP	1.29	268	80.2	54.9	100
TDCPP	0.19	44.5	4.34	2.57	100
TEP	0	350	43.0	13.4	93
TiBP	0.18	218	13.4	3.67	100
TBP	0.10	80.9	6.27	1.59	100
TBEP	0	47.2	4.21	1.17	73
THP	0	104	3.47	0.00	50
TPP	0	15.7	0.95	0.34	98
TiPrP	0	14.7	0.70	0.09	70
TPrP	0	5.06	0.63	0.20	75
TCP	0	15.0	0.39	0.00	95
TEHP	0	3.33	0.35	0.14	78
TPeP	0	3.07	0.17	0.02	58
TMP	0	1.63	0.08	0.00	5
TPPO	0.70	5850	224	19.8	100

Table 3. Concentrations of selected OPEs and TPPO in different studies in river water in the world (ng L⁻¹)

River	Location	TCCP	TCEP	TEP	TiBP	TDCPP	TBP	TPP	TBEP	TPPO	Ref.
40 rivers	North China	5-921	1-268	<LOD-350	<LOD-218	<LOD-44	<LOD -81	<LOD-16	<LOD-47	<LOD -5852	This study
Pearl River Estuaries	South China	150-1150	220-1160	n.a.	n.a.	n.a.	n.a.	n.a.	n.a.	n.a.	9
Songhua River	Northeast China	5-190	38-3700	5-190	n.a.	2-46	87-960	5-65	5-310	n.a.	23
Elbe	Germany	40-250	5-20	10-180	10-50	n.a.	2-8	<LOD-4	<LOD-80	10-40	1
Rhine	Germany	75-160	12-25	n.a.	17-84	n.a.	6-28	1-2	28-54	n.a.	1
Four streams	Germany	<LOD-2914	<LOD-557	n.a.	n.a.	<LOD-1284	<LOD-3889	n.a.	<LOD-1773	n.a.	22
Aire	UK	113-26050	119-316	n.a.	n.a.	62-149	n.a.	6-22	n.a.	n.a.	31
Aire	UK	4821	181	n.a.	n.a.	49	n.a.	17	n.a.	n.a.	32
Navarra, Asturias, Catalonia	Spain	<LOD-1800	<LOD-330	n.a.	<LOD-1200	<LOD-200	<LOD-370	<LOD-18	<LOD-4600	n.a.	33
Three rivers	Austria	33-170	13-130	13-51	n.a.	<LOD-19	20-110	<LOD-10	24-500	n.a.	34
three major rivers	South Korea.	n.a.	42	n.a.	n.a.	n.a.	n.a.	n.a.	n.a.	n.a.	35
Arkansas streams	USA	n.a.	48-700	n.a.	n.a.	n.a.	31-560	n.a.	n.a.	n.a.	36

<LOD: below limit of detection; <LOQ: below limit of quantification; n.a.: not available